On the influence of sewage pollution on inshore benthic communities in the South of Kiel Bay

Part 2. Quantitative studies on community structure

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KURZFASSUNG: Zum Einfluß kommunaler Abwässer auf küstennahe benthische Gemeinschaften in der südlichen Kieler Bucht. Teil 2. Quantitative Untersuchungen zur Struktur der Gemeinschaften. Innerhalb eines Verschmutzungsgradienten von 2000 m Radius um die Einleitungsstelle kommunaler Abwässer bei Bülk (westliche Ostsee) konnten auf Sandböden drei Assoziationen gefunden werden. Sie sind charakterisiert durch: (a) den Polychaeten *Capitella capitata* und Oligochaeten (50–100 m Entfernung von der Einleitungsstelle der Abwässer); (b) den Polychaeten *Pygospio elegans* (200–ca. 700 m) und (c) den Amphipoden *Bathyporeia sarsi* (> 700 m). Die Assoziationsgrenzen sind verschwommen, es liegt ein Kontinuum vor. Der Einfluß biogener Substratstrukturen wird im schwächer verschmuzten Bereich deutlicher. Hier können Assoziationen klar abgegrenzt werden. Die Kombination der Lebensformtypen, Artenzahl, Diversität und andere Eigenschaften der Gemeinschaften hängen in charakteristischer Weise von der räumlichen Heterogenität des Substrates ab. Sie werden ebenfalls stark vom Verschmutzungsgrad beeinflußt. Die küstennahe benthische Makrofauna liefert einen wichtigen Beitrag zur Selbstreinigung des ganzen Ökosystems, indem es partikuläre organische Substanzen zu nutzbarer Fischnahrung umformt. Abundanz und Biomasse werden dabei vervielfacht. Dieser Vorgang bringt zusätzlich zum Eutrophierungsproblem die Gefahr der Anreicherung von Giftstoffen mit sich. Komplexere Systeme (z. B. Miesmuschelbänke) sind besser auf solche Faktoren eingestellt, die unmittelbar die Äquität der Gemeinschaft beeinflussen. Der ökologische Puffer besteht aus vielen spezialisierten Konkurrenten und Räubern. Niedrig diverse Systeme (z. B. Sandböden) reagieren mit hohen Amplituden auf Störungen dieses Typs. Zusammensetzung, Aquität, Artenreichtum, Abundanz und Biomasse tierischer Gemeinschaften sind brauchbare Parameter einer ökologischen Methode zur Früherkennung von Umweltbelastungen.

INTRODUCTION

For several decades, particularly in the last few years, an increasing endangering of the Baltic Sea through spreading of abiotic areas characterized by oxygen insufficiency and hydrogen sulfide has been registered (FONSELIUS, 1962, 67, 69; ICES report, 1970). The main causes are probably a steady increase of salinity (FONSELIUS, 1969) together with temperature (KULLENBERG, 1970). Pollution caused by all countries bordering on the Baltic Sea has also contributed to the problem.

In this situation facts and models are needed concerning the reactions of different Baltic Sea areas and of their ecosystem fractions on all kinds of disturbances, whether of natural or human origin. This paper deals with an inshore benthic subsystem partially polluted by sewage, which was described in a preliminary report after a half year pilot study (ANGER, 1975). Particular attention is paid to seasonal and pollution-influenced changes of the community structure, on production in different levels of benthic eutrophication (in the widest sense), and to the role of the substrate type. The relationships between the system's "spatial heterogeneity", "physical stability", "complexity" (measured as diversity), and "biotic constancy" were observed. The question is raised whether an ecological structure analysis as performed here is an appropriate method of measuring environmental stress.

THE INVESTIGATION AREA

In the South of Kiel Bay near Bülk about 50,000 m^3 of untreated sewage was pumped every day into the Baltic Sea. The investigation area was thoroughly described by HORSTMANN (1972), HUBRICH (1972) and ANGER (1975). The bacterial contamination due to the sewage pollution was investigated by GRAMM (1936), RHEINHEIMER (1966, 67), GÄRTNER (1969) and SADJEDI (1971). Hydrographical studies were performed in Kiel Bay by WATTENBERG (1949), KÄNDLER (1952, 59, 60), OHL (1959), KRUG (1963) and other workers. HORSTMANN (1972) elaborated data on plankton and discussed older literature. A model of the dispersion of the pollutants was calculated by HUBRICH (1972). Some information on phosphate, silicate, seston and other pollution parameters is given by HORSTMANN (1972) and ANGER (1975).

Table 1

Quality and composition of the sewage discharged at Bülk

pH	7.55 $\pm$	0.18	
Precipitate	9.1 $\pm$	4.6	ml/l
BOD$_5$	225 $\pm$	80	mg/l
KMnO$_4$-consumption (unfiltered)	568 $\pm$	183	mg/l
Decomposition capability (Methylene Blue samples)	3.2 $\pm$	1.6	hours
O$_2$-content	0 $\pm$	0	mg/l
Ammonium (NH$_4^+$)	21.4 $\pm$	7.2	mg/l
Ammonium-nitrogen	16.7 $\pm$	5.6	mg/l
Organic fixed nitrogen	11.4 $\pm$	2.5	mg/l
Total nitrogen	28.1 $\pm$	5.8	mg/l
Phosphates	48 $\pm$	11.7	mg/l
Chlorides	138.4 $\pm$	17.8	mg/l
Evaporation residue (110° C)	938 $\pm$	128	mg/l
Ignition residue (600–650° C)	575 $\pm$	97	mg/l
Ignition loss (600–650° C)	363 $\pm$	49	mg/l
	$\triangleq$ 38.9 $\pm$	3.8	%

Table 1 presents additional data on the composition and some other attributes of the sewage. It consisted mainly of organic particles; poisonous substances do not play

any apparent role. The sampling cruises concentrated on the 3 m depth line. Here there is a medium fine sand with a median grain size of 250–300 μ, which had to be regarded as "less well sorted" (sorting coefficient: 50–200 φ). The ignition loss showed significantly higher values up to more than 7 % in the heavily polluted zone, but just a slight increase in the transition area (200–500 m distance from the sewage outlet) and a normal level (below 1 %) in all other stations. For further details of the sediment quality and its meiobenthos see ANGER & SCHEIBEL (in press).

In contrast to the grain size some macroorganisms had a high influence on the benthic populations: Mussel beds *(Mytilus edulis)* and eel-grass *(Zostera marina)* altered the substrate conditions completely and therefore they were regarded as dominant factors in this study.

The oxygen saturation was measured in the near bottom water by means of the TO-probe described by GRASHOFF (1969). Since all values averaged about 100 %, oxygen was considered to be a less important factor for the macrobenthos. The salinity in the investigation area normally fluctuates between 13 and 18 %. It has no importance in this case, as neither horizontal nor vertical gradients could be observed. The water temperature was recorded every day and related to spawning periods of benthic animals (for details see ANGER, in press).

Thus changes in the composition of faunal assemblages can be caused by seasonal temperature fluctuations, differing concentrations of sedimenting organic particles, or dominant substrates. Other factors are less significant.

MATERIALS AND METHODS

Nine inshore stations were investigated at constant distances from the sewage outlet: 50, 100, 200, 300, 500, 700, 1000, 1500, 2000 m. These distances are used as measures for the degree of sewage pollution; for data on the dilution rates see HORST-MANN (1972), HUBRICH (1972) and ANGER (1975). From August 1971 until September 1972 129 macrofauna samples were taken by means of a 0.1 m² van Veen grab. In addition 28 samples collected in June and July 1971 from other stations were considered for some comparisons.

Each sample consisted of two parallel grabs and thus represented 0.2 m². Its content was immediately strained through a sieve with 1 mm mesh size and preserved in a 4 % formaline solution. In order to investigate young bottom stages of macrobenthic animals during the year simultaneous meiofauna samples were taken at each pure sand station (cf. ANGER & SCHEIBEL, in press). The material was sorted and identified under a stereo microscope in the laboratory. The biomass was determined as dry weight of samples dried at 100° C to constant weight. For the calculation of diversity indices a PDP 10 computer belonging to Kiel University was used.

RESULTS

General Remarks

Table 2 gives a rough idea of the macrobenthic colonization of the whole investigation area. The 34 most common species are ranked according to the number of station samples in which they occurred. The percentage frequency or presence (P) and mean dominance (= individual percentage of a certain species = $\bar{D}$) is indicated for the main substrate types, sand (S), sea grass (Sg) and mussel beds (M). This allows the recognition of substrate preferences.

Table 2

List and rank of the most common species according to the number of stations (St.) they were found. Presence (P) and mean dominance ($\bar{D}$) on sand (S) seagrass (Sg), and mussel beds (M)

Species	St.	P %			$\bar{D}$ %		
		S	Sg	M	S	Sg	M
Capitella capitata	146	89.5	95.2	100	19.4	1.6	9.8
Macoma baltica	145	87.7	100	100	3.6	2.4	0.8
Hydrobia ulvae	141	88.9	96.2	89.5	6.0	6.0	2.4
Mytilus edulis	140	84.8	96.2	100	2.4	5.7	29.2
Pygospio elegans	136	91.2	100	100	26.1	42.5	8.3
Nereis diversicolor	128	84.2	66.7	81.3	2.9	0.5	0.6
Mya arenaria	126	73.7	88.5	94.7	4.2	5.8	2.8
Eteone longa	104	66.7	100	63.2	0.7	1.1	0.4
Corophium insidiosum	100	28.2	100	90.0	0.5	5.3	9.4
Gammarus salinus	88	36.8	57.1	87.5	0.5	0.1	2.7
Anaitides mucosa	84	45.6	95.2	62.5	0.2	0.6	0.1
Idotea baltica	83	32.3	96.2	94.7	—	1.9	1.1
Microdeutopus gryllotalpa	81	30.3	92.3	89.5	0.1	1.7	4.7
Cardium lamarcki	77	40.6	87.0	55.6	5.6	1.2	0.1
Scoloplos armiger	75	54.4	85.7	56.3	0.9	3.2	1.8
Arenicola marina	73	43.9	61.9	56.3	0.5	0.2	0.4
Littorina littorea	72	39.1	69.6	77.8	0.2	0.4	0.5
Spio filicornis	69	71.9	38.1	18.8	4.4	0.1	0.3
Bathyporeia sarsi	69	87.0	27.8	28.6	10.0	0.1	0.7
Polydora ciliata	68	22.2	84.6	89.5	0.2	2.7	5.5
Polydora ligni	64	24.6	57.1	68.8	0.1	0.5	1.7
Crangon crangon	51	42.1	66.7	25.0	0.3	0.1	—
Asterias rubens	49	11.6	82.6	83.3	—	0.6	0.6
Gammarus oceanicus	48	10.1	65.4	73.7	—	0.2	0.8
Harmothoe impar	44	15.8	38.1	100	—	—	0.6
Nereis succinea	41	13.1	42.3	78.9	0.1	0.1	0.7
Calliopius laeviusculus	37	11.1	50.0	42.1	—	0.1	0.1
Fabricia sabella	35	8.8	61.9	56.3	—	5.0	0.3
Lacuna divaricata	34	8.1	61.5	52.6	—	0.3	0.2
Balanus crenatus	34	8.1	23.1	78.9	—	—	1.3
Pholoe minuta	33	1.8	57.1	68.8	—	0.2	0.2
Jaera albifrons	33	10.1	43.5	72.2	—	0.1	0.2
Carcinus maenas	32	10.5	57.1	75.0	—	0.1	0.2
Rissoa inconspicua	25	7.1	42.3	42.1	—	0.1	0.1

Pollution and seasonal influenced changes

The annual mean biomass $(\overline{B})$[*] and abundance $(\overline{A})$ of the entire macrofauna is increased on average by a factor of 10 in the semipolluted area as compared with unaffected stations (Fig. 1).

Three zones can be distinguished within the investigation area: Zone I: 50–200 m distance from the outlet, zone II: 300–700 m distance from the outlet, zone III: 1000–2000 m distance from the outlet.
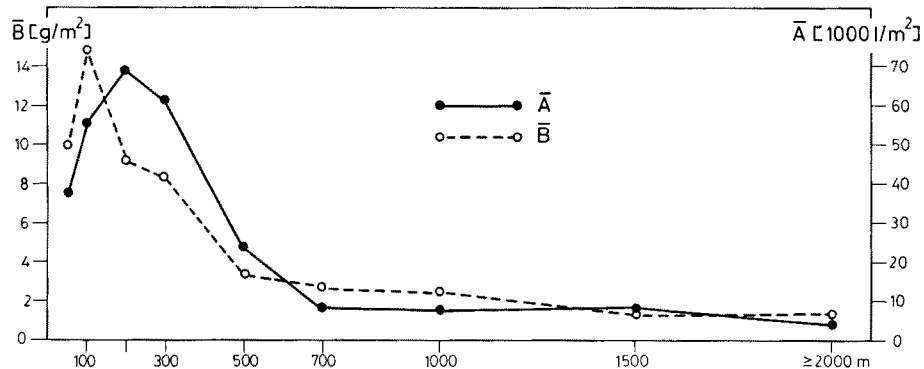

Fig. 1: Mean abundance $(\overline{A})$ and biomass $(\overline{B})$ within the pollution gradient

When these zones are compared in terms of an annual cycle of abundance and biomass (Fig. 2) and of the percentage of major groups (Fig. 3), significant differences are apparent: Polychaetes, which are nearly always dominant (mainly *Capitella capitata*), and oligochaetes both increase their individual numbers by high propagation rates in zone I during July.

In the less polluted subareas II and III the same picture appears somewhat later due to settling of mulluscan larvae *(Mya arenaria, Macoma baltica, Cardium lamarcki, Hydrobia ulvae)*, while at the same time in zone I the young annelid populations are greatly reduced. In autumn in the sewage affected zones I and II a new population rise takes place, this is attributed to the polychaete *Pygospio elegans*. In the nearly unaffected part III the relative importance of the amphipod *Bathyporeia sarsi* increases at the same time. After the general decrease in population density at beginning of winter a new peak occurs in February in zones II and III. It can be attributed partly to *Pygospio elegans* (II), and partly to several amphipod and polychaete species (III).

In zone III the abundance values remain constantly low until July, while the corresponding biomass rises. During this time the dominating worms *Pygospio elegans*

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[*] To avoid a masking of trends by irregularly occurring animals with heavy calcareous bodies, all representatives of Mollusca, Echinodermata, Cirripedia and *Carcinus maenas* were excluded from biomass calculations.

and *Capitella capitata* produce high population densities again. The culmination of the polychaete propagation phase is finally reached in July.

At this time *Capitella capitata* clearly dominates zone I. *Pygospio elegans* which is also very abundant here, dominates zone II even more conspicuously.

In summer 1972 the biological composition of the three zones was remarkably similar. This fact can be interpreted as follows: Simultaneously with the annelids men-
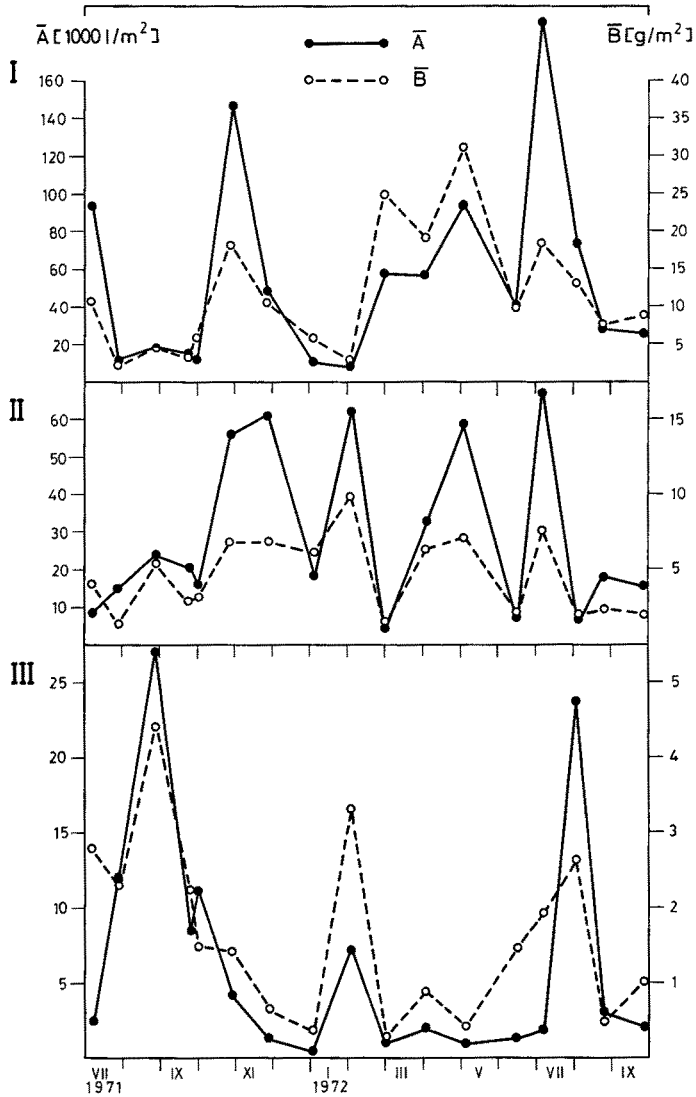

Fig. 2: Mean abundance (Ā) and biomass (B̄) during the year (I = 50–200 m, II = 300–700 m, III = 1000–2000 m distance from the outlet)
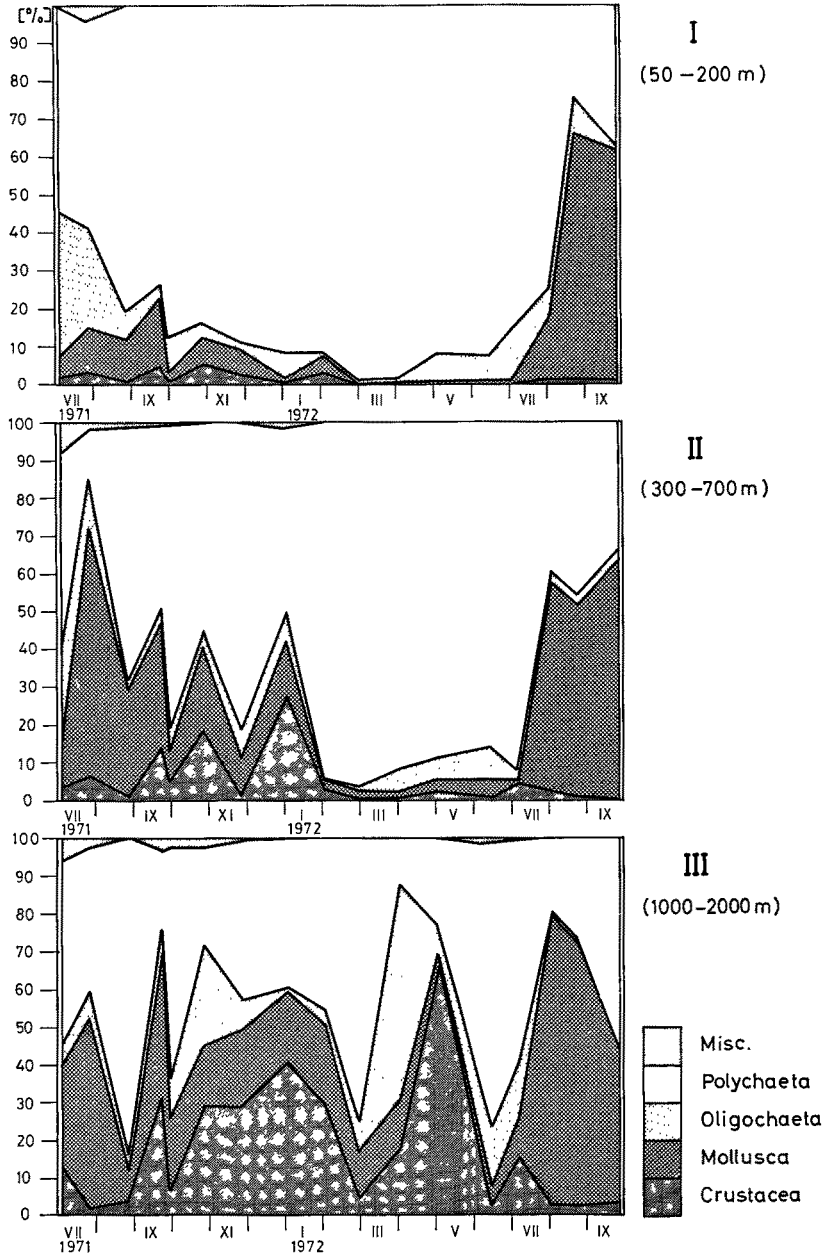
Fig. 3: Percentage of the most important macrofauna groups during the year

tioned above, the predator *Nereis diversicolor* also had an unusually strong reproductive phase, when the mechanical sewage treatment was started. Thus a considerable abatement of food supply and contemporary predation by normally omnivorous animals reduced the population densities of the typical detritus feeding polychaetes *Capitella capitata* and *Pygospio elegans*. During this period masses of molluscan larvae settled on the sandy bottom and appeared as additional competitors.

It must be assumed that later the faunal assemblages of zones I and II were altered remarkably, becoming more and more similar to those of zone III. This development was not investigated.

Interspecific association

Community analysis was introduced by evaluating species groups which are connected by a high degree of interspecific association. For this purpose every possible species pair was examined for its distribution affinity using the index proposed by JACCARD (1912) for phytosociological studies.

It is calculated by the formula

$$J = c / (a + b - c)$$

where a = number of stations (samples) containing species A, b = number of stations (samples) containing species B, c = number of stations (samples) containing species A and B together.

After transforming to percentage values ("percentage affinity") the matrix was arranged in a Trellis diagram. It was started with the species showing the highest number of index values over 50, followed by the species with the highest affinity to the first one. The succeeding rank resulted from the respective highest affinities to the already existing group. The whole procedure was carried out separately for the regions 50–500 m and 700–2000 m distance from the sewage pipeline.

Figure 4 shows the association pattern in the rather unaffected zone. The most conspicuous group consists of the commonest species inhabiting sand flats and eel grass meadows (cf. Table 2): *Mytilus edulis, Macoma baltica, Pygospio elegans, Hydrobia ulvae, Eteone longa, Capitella capitata, Mya arenaria, Nereis diversicolor, Cardium lamarcki, Scoloplos armiger, Arenicola marina, Spio filicornis, Bathyporeia sarsi.*

A second accumulation of high indices appears in the left upper corner of the diagram. It is composed of animals living in phytal and other growth: *Corophium insidiosum, Idotea baltica, Asterias rubens, Littorina littorea, Polydora ciliata, Anaitides mucosa, Harmothoe impar, Polydora ligni, Pholoe minuta, Fabricia sabella, Gammarus oceanicus, Jaera albifrons, Gammarus salinus, Lacuna divaricata.*

The rest of the species are too sparse to reach high Jaccard-values. *Crangon crangon* has a conspicuous affinity to the first group, the other species tend to the second one: *Nereis succinea, Rissoa inconspicua, Carcinus maenas, Calliopius laeviusculus, Balanus crenatus.*

In general, a strong influence of biotic substrate structures on macrofaunal association is visible.

In the polluted zone (50–500 m) the affinity pattern is quite different (Fig. 5). The seven commonest macrobenthic species are extremely strongly associated. All are more or less favoured by organic pollution and under well defined preconditions they can
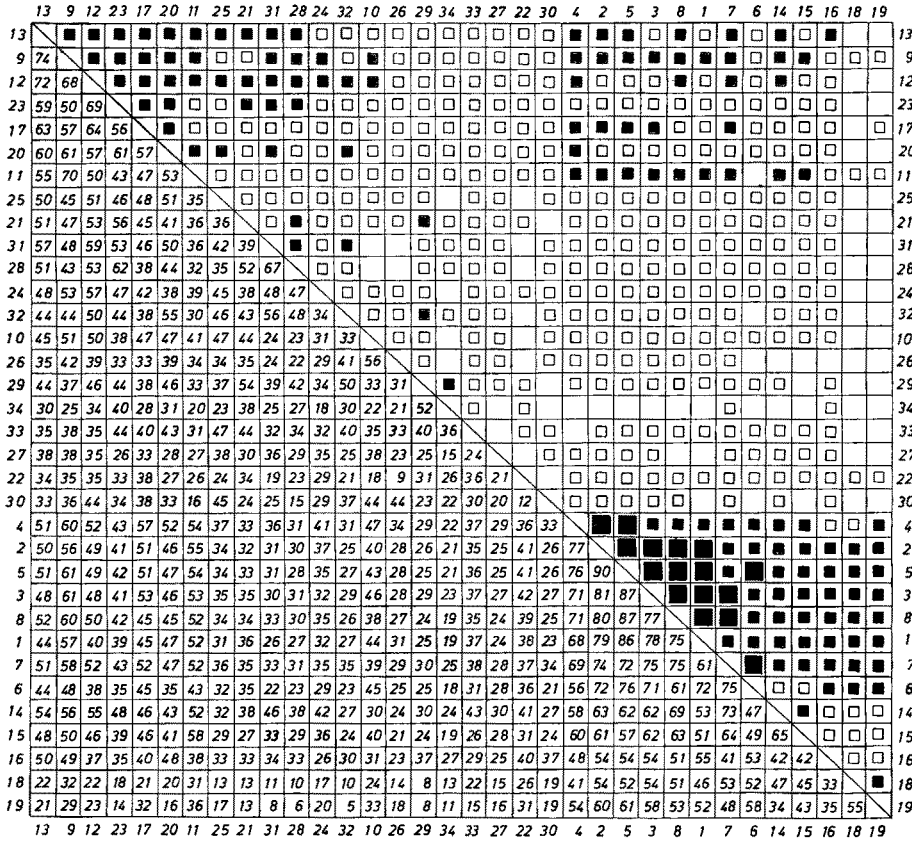
Fig. 4: Interspecific association, expressed as percentage affinity (JACCARD-index × 100). Stations from 700 to 2000 m distance from the outlet. For numbers of species see Table 2

have indicator quality: *Pygospio elegans, Mya arenaria, Macoma baltica, Hydrobia ulvae, Nereis diversicolor, Mytilus edulis, Capitella capitata.*

Seven euryoeceous growth dwellers with some indicator value show a high affinity to this group and complete it to a typical recurrent "indicator association" for organic pollution: *Microdeutopus gryllotalpa, Corophium insidiosum, Anaitides mucosa, Idotea baltica, Gammarus salinus, Eteone longa, Polydora ligni.*

A further group is made up of some species with only weak affinity to the stricter indicator group, but strong relationship to the last mentioned seven species: *Cardium lamarcki, Polydora ciliata, Scoloplos armiger, Littorina littorea, Gammarus oceanicus, Arenicola marina, Asterias rubens, Nereis succinea, Crangon crangon, Harmothoe impar, Spio filicornis, Bathyporeia sarsi, Calliopius laeviusculus.*

The rarest species almost never occur in the polluted zone together with the indicators, but show mutually a higher affinity: *Fabricia sabella, Lacuna divaricata, Jaera albifrons, Balanus crenatus, Pholoe minuta, Rissoa inconspicua, Carcinus meanas.*
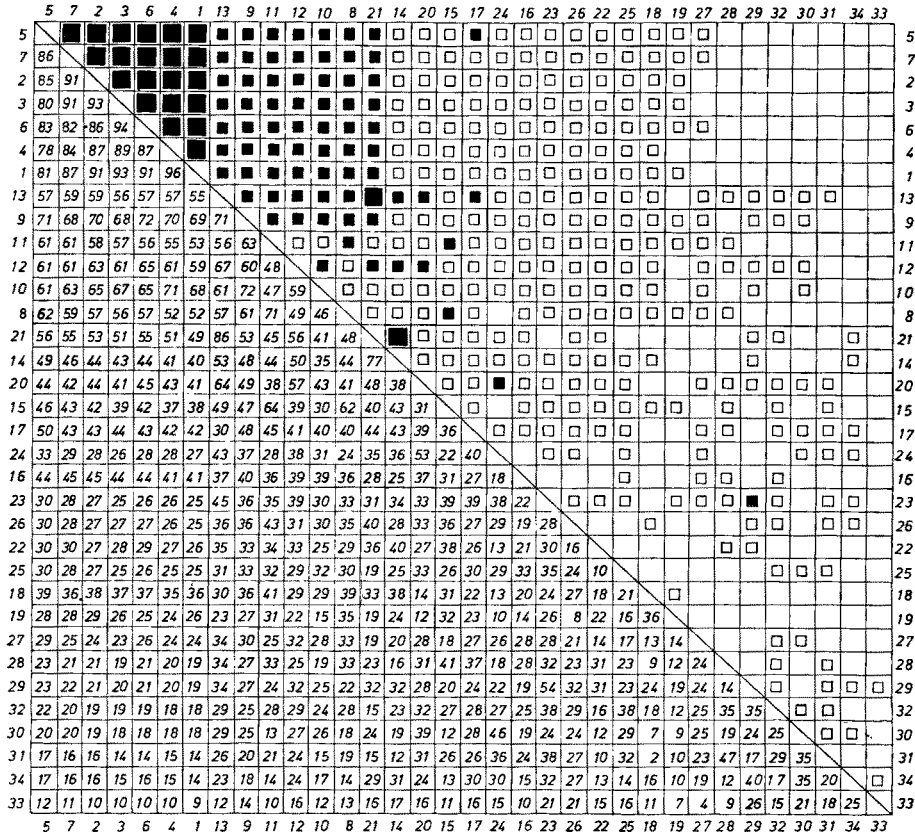

Fig. 5: Interspecific association (cf. Fig. 4). Stations from 50 to 500 m distance from the outlet

The pattern in the sewage affected reaches is primarily physically controlled. Species with a wide ecological range dominate clearly; biotic substrate properties do not have the critical importance they usually have.

Substrate specifity

To get a more refined picture from the ecological groupings one more approach is necessary. It is based on the more or less pronounced substrate preference of the single species. This can be expressed by relativating of population parameters as percentage frequency (P) and mean abundance ($\overline{A}$):

$$P_r = \frac{P \times 100}{P_\Sigma} \quad (\%) \qquad\qquad \overline{A}_r = \frac{\overline{A} \times 100}{\overline{A}_\Sigma} \quad (\%)$$

where P_r, $\overline{A}_r$ = relative presence, relative mean abundance; P, $\overline{A}$ = mean values, observed on a certain substrate; P_Σ, $\overline{A}_\Sigma$ = sum of the P or A on all three distinguished substrates together.

The P_Σ, $\overline{A}_\Sigma$ values correspond to 100 %. The P_r and $\overline{A}_r$ values calculated in this way were arranged in triangular coordinate systems, since they always add up to 100 % on all three substrates. To avoid overlapping influences of heavy pollution effects, all data from zone I were excluded from the calculations.
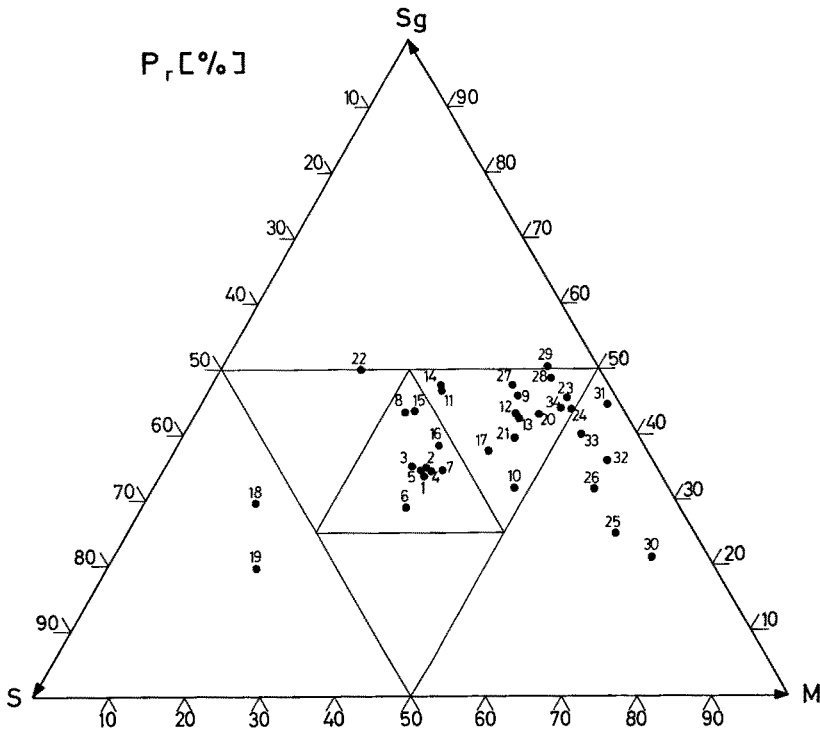

Fig. 6: Substrate specifity of the most common macrofauna species according to their relative frequency (P_r). For abbreviations and numbers of species see Table 2

It must be stressed that the position of a species within the coordinate system is only an expression of its substrate specifity, not of its importance for the corresponding community.

The P_r values (Fig. 6) make a group of eurytopous species visible, concentrated in the small central triangle. They are nearly identical with the "indicator association" identified by the Trellis diagram. They occur on the three substrates with about the

same frequency. In the overlapping triangle between sea grass (Sg) and mussel beds (M) there is a group of typical aufwuchs dwelling species, which are not restricted to one single substrate, but occur much less frequently on pure sand.

Crangon crangon tends to *Zostera* biotopes, but has also strong relationships with vegetation-free sand. Only a few species are already restricted to mussel beds by their frequency of occurrence, two *(Spio filicornis* and *Bathyporeia sarsi)* to sand bottoms, and just one *(Lacuna divaricata)* to eel grass.

According to Figure 7 the $\bar{A}_r$ values give a quite different pattern: No one species is completely indifferent regarding its quantitative distribution. Only a few species are nearly equally abundant on sand and *Zostera* meadows and only one is equally abun-
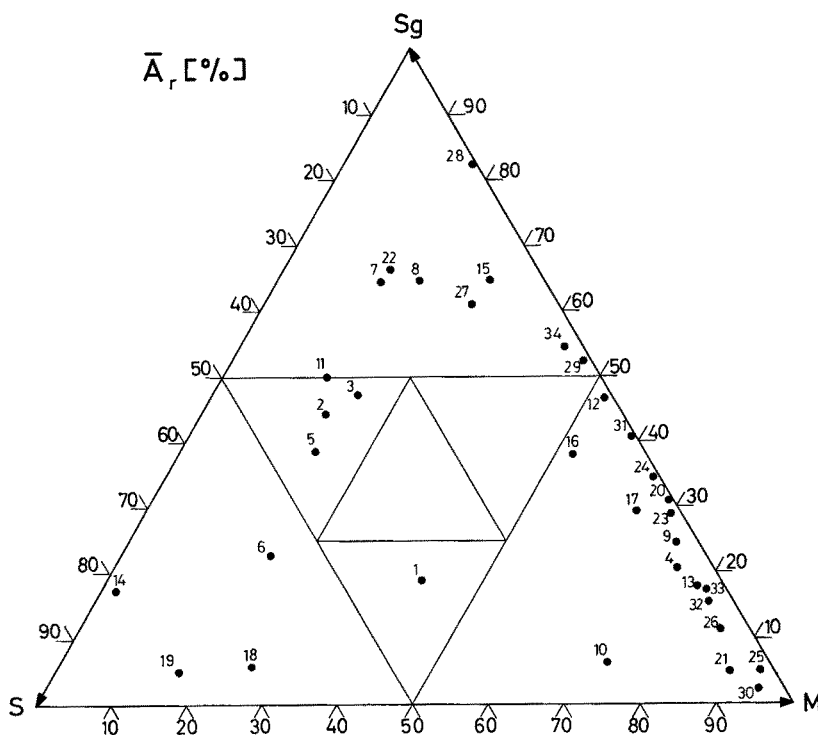

Fig. 7: Substrate specifity of the most common macrofauna species according to their relative mean abundance ($\bar{A}_r$) (cf. Fig. 6)

dant on mussel beds and sand *(Capitella capitata)*. All other macrobenthic species show a more or less clear quantitative preference for one single substrate type.

In the following chapters the main benthic subsystems of the investigation area are biologically characterized by grouping species by substrate specifity and by ranking species according to their $\bar{A}P^{1/2}$-values. These are an expression of the quantitative importance for a community (ERMAN & HELM, 1971).

The sand bottom

The majority of the sample material originates from more or less pure sand areas. Thus the most information comes from this simplest substrate type.

Table 3

Sand bottom community: Species grouped according to their substrate specifity. Explanations: a = specific as regards both relative frequency and abundance; b = only specific as regards relative abundance; c = more frequent and / or abundant on two substrates; d = indifferent as regards frequency, more abundant on other substrates

a	b	c	d
Bathyporeia sarsi	Cardium lamarcki	Crangon crangon	Mytilus edulis
Spio filicornis	Nereis diversicolor	Pygospio elegans	Eteone longa
		Capitella capitata	Mya arenaria
		Macoma baltica	Scoloplos armiger
		Hydrobia ulvae	Arenicola marina
		Anaitides mucosa	

As can be established from Figures 6 and 7 only a few species are characteristic for this subsystem (group "a" in Table 3). They can be supplemented by the highly specific, but sparsely occurring *Pseudocuma longicorne* (Cumacea) and *Bathyporeia pilosa* (Amphipoda).

A completely different rank can be set up regarding the quantitative importance (measured as $\overline{A}P^{1/2}$). Table 4 is completed by the systematic groups Oligochaeta and Nemertini. Besides the index $\overline{A}P^{1/2}$ the following values are given: mean and maximum abundance ($\overline{A}$, A_{max}), mean abundance in the heavily polluted zone ($\overline{A}_{50\text{-}300\,m}$), mean and maximum biomass ($\overline{B}$, B_{max}), mean and maximum dominance ($\overline{D}$, D_{max}), and frequency (P). All values except $\overline{A}_{50\text{-}300\,m}$ refer to the slightly affected and unpolluted zones only.

Table 4 indicates that euryoeceous species (cf. Figs 4–7) dominate the sand community, while characteristic animals have a medium or lower place in the rank.

Life forms

Sand gives the best possibilities for expansion for deposit feeders. They form the dominating group with 58 % of the species and 90 % of the individuals. Predators, with 26 % of the species and just 1 % of the individuals and suspension feeders with 16 % of the species and 9 % of the individuals are less important. The commonest (euryoeceous) selective deposit feeders *(Capitella capitata, Pygospio elegans)*, together with those suspension feeders capable of ingesting coarser particles *(Mya arenaria, Mytilus edulis)* and very unspecialized predators *(Nereis diversicolor)*, are favoured by sedimenting organic substances. Thus it can be stated that the sandy bottom macrofauna is preadapted to transform a high amount of sewage particles to valuable biomass for the food web (cf. Fig. 9). It must be assumed that it carries out a large portion of the self purification of the whole ecosystem of Kiel Bay.

Table 4

Sand bottom community: rank of species according to their ĀP^½-values
(explanations see text)

Species	ĀP^½	Ā[I/m²] 50–300 m	Ã[I/m²]	A[I/m²] max	B̄[mg/m²]	B[mg/m²] max	D̄[%]	D[%] max	P[%]
Pygospio elegans	11396	11934	22628	147500	1964,8	15045,0	26,1	87,3	91,2
Capitella capitata	1624	1717	21054	29195	312,9	4817,2	19,4	97,1	89,5
Hydrobia ulvae	1006	1067	1608	13710	798,6	11790,6	6,0	37,6	88,9
Mya arenaria	522	608	915	9420	3240,4	35268,3	4,2	45,9	73,7
Cardium lamarcki	375	589	345	10325	476,9	10951,2	5,6	69,3	40,6
Mytilus edulis	332	360	325	3405	5530,0	69850,0	2,4	47,6	84,8
Macoma baltica	275	293	289	7260	4041,1	30111,5	3,6	16,5	87,8
Nereis diversicolor	159	173	700	1873	472,0	2591,8	2,9	16,1	84,2
Oligochaeta	129	155	4341	11805	21,1	1463,8	6,5	61,6	69,2
Bathyporeia sarsi	96	103	11	820	50,1	255,5	10,0	65,3	87,0
Spio filicornis	70	82	21	1230	22,4	243,5	4,4	51,0	71,9
Scoloplos armiger	32	44	14	1185	213,8	4429,5	0,9	23,1	54,4
Anaitides mucosa	20	29	34	370	49,0	886,0	0,2	1,8	45,6
Bathyporeia pilosa	17	20	0	800	9,3	345,5	2,0	64,8	30,4
Etone longa	16	20	10	185	28,6	207,5	0,7	7,1	66,7
Arenicola marina	5	7	5	55	172,3	3324,0	0,5	10,2	43,9
Nemertini	4	6	3	35	18,9	199,5	0,2	3,0	36,8
Crangon crangon	3	4	2	20	13,6	430,5	0,3	6,1	42,1
Pseudocuma longicorne	1	2	0	50	—	—	0,1	2,1	14,5

Most of the sand bottom species (53 %) and 80 % of the individuals belong to the hemisessile infauna. Motile animals, mostly predators, are much less numerous, but provide relatively many species.

Faunal composition in the pollution gradient

Figure 8 shows the strong influence of sewage pollution on the composition of the macrobenthic fauna (annual mean). Within a radius of about 100 m *Capitella capitata* is clearly dominant. From 200 to 700 m it becomes less important compared with *Pygospio elegans*. Beyond this zone the diversity becomes greater, miscellaneous species become more prominent in the community; *Bathyporeia sarsi* becomes more frequent and numerous, indicating a rather clean sand.
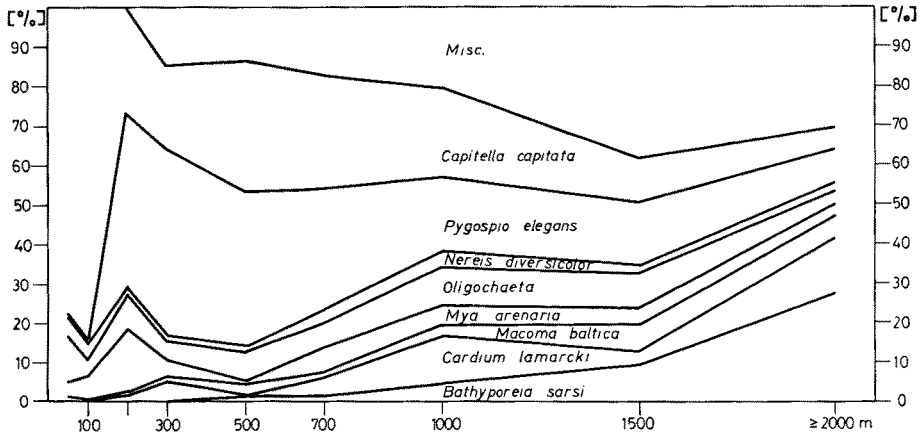

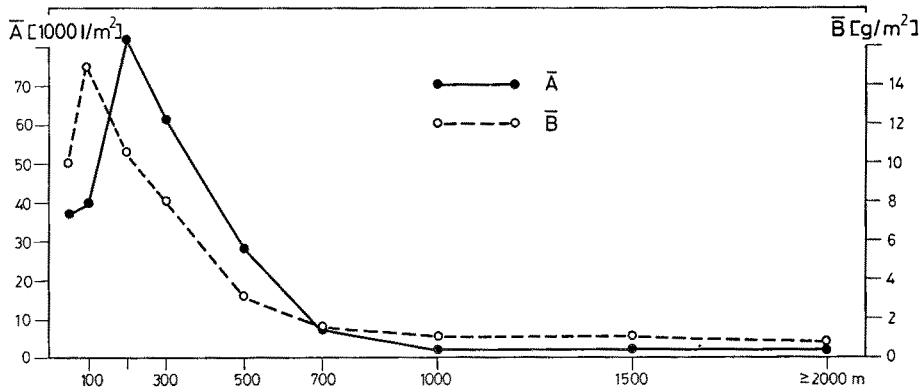Fig. 8: Macrofauna composition in the pollution gradient (sand bottom)

Fig. 9: Sand bottom macrofauna: mean abundance (A̅) and biomass (B̅) in the pollution gradient

As stated above, correlated with feeding types, a remarkably great rise of production occurs in the polluted zone of the sand flats. It appears as a multiplication of mean biomass by a factor of 15 and of mean abundance by a factor of 50 (Fig. 9). Thus it exceeds the average considerably (cf. Fig. 1). This fact underlines again the eminent importance of the sand dwelling inshore macrofauna for the self-purification of the whole ecosystem.

Nevertheless one must take into consideration the fact that a high transformation of sewage to biomass is not positive per se, since it is linked with an equally high accumulation of toxic substances which also occur in sewage from municipalities. The second danger to be borne in mind is still largely unknown in its consequences for the ecosystem: the decline of species diversity.

There are at least two reasons for this pauperization. The primary one is a direct stress situation caused by sedimenting particles and cellulose fibres contaminating and clogging the habitat niches, and filter and gill apparatus. This must account for the fact that *Capitella capitata* and all other species non-selective for habitat and food show a conspicuous decrease of population densities in the severely polluted zone (50–100 m radius; see Figs. 1 and 9). Mainly in the semipolluted zone a second effect is important: With the increase in food supply by sewage particles, euryoeceous, preadapted populations compete more successfully for space in the biotope. More sensitive and specialized species are repressed in two ways under those conditions.

Figure 10 shows the behaviour of diversity and species number in the pollution gradient. In order to compare some current diversity indices and to ascertain the trend disussed above the following measures were used:

$$(1) \quad H' = - \sum_{i=1}^{S} \frac{N_i}{N} \, \mathrm{ld} \left(\frac{N_i}{N} \right) \quad \text{(SHANNON \& WEAVER, 1949; cf. also PIELOU, 1966a, b, 1969)}$$

$$(2) \quad F = 1 - \sum_{i=1}^{S} \frac{N_i \, (N_i - 1)}{N \, (N - 1)} \quad \text{(SIMPSON, 1949; PIELOU, 1969)}$$

$$(3) \quad D_s = 1 / \sum_{i=1}^{S} \frac{N_i \, (N_i - 1)}{N \, (N - 1)} \quad \text{(SIMPSON, 1949; McARTHUR \& WILSON, 1967; McARTHUR, 1972).}$$

The evenness component of species diversity was calculated as

$$J' = \frac{H'}{\mathrm{ld} \, S} \quad \text{(PIELOU, 1966a, b, 1969)}$$

The observations on the faunal compositions indicate a disturbance of the subsystem as a consequence of pollution stress. This is confirmed by all measures of diversity and evenness (Fig. 10). The curves are inversely correlated with the mean abundance and biomass (Fig. 9), and also with the dilution and ignition loss values. The mean number of macrobenthic species ($\bar{S}$) shows a diverging trend: There is a slight maximum in the semipolluted zone around 300 m distance from the outlet. This apparant anomaly can be explained by a sort of "edge effect". Several euryoeceous aufwuchs dwellers, which normally do not occur in pure sand, live frequently in organically enriched sediment (e. g. *Corophium insidiosum, Microdeutopus gryllotalpa, Ido-*
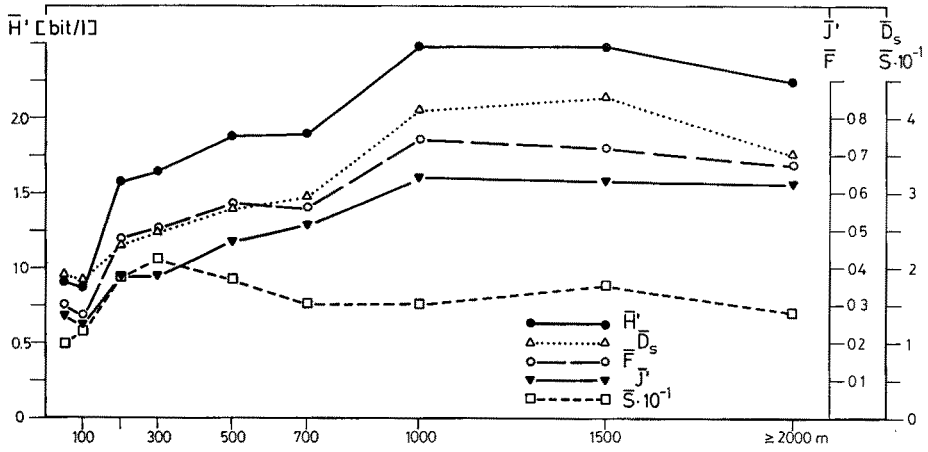
K. ANGER

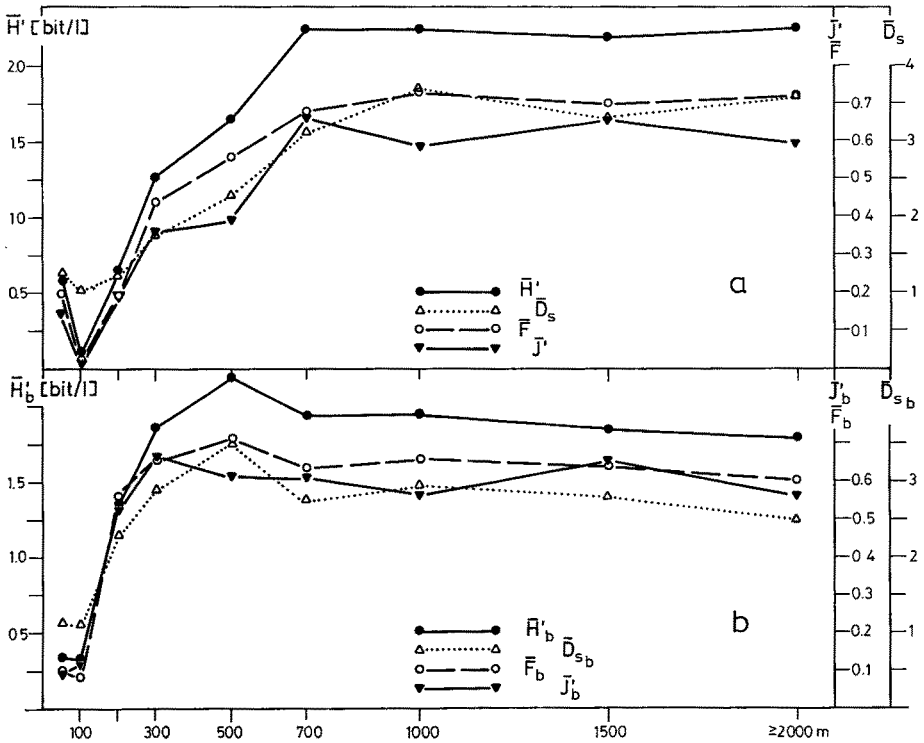Fig. 10: Sand bottom macrofauna: mean number of species ($\overline{S}$) and diversity in the pollution gradient

Fig. 11. Sand bottom macrofauna: comparison between numerical (a) and biomass (b) diversity in the pollution gradient

tea baltica, Gammarus salinus, Gammarus oceanicus). Here the sand is pure enough to allow the survival of sensitive species and sufficiently enriched to harbour animals of the *Capitella*-type (*Capitella capitata*, Naididae, Tubificidae).

Thus is can be formulated that in a sand bottom subsystem, environmental stress mainly affects the evenness component. This agrees well with the hypothesis of TRAMER (1969) that under unpredictable, in some respect extreme, conditions diversity is regulated by the evenness, while a predictable (stable) system is regulated by the species richness.

The diversity values are negatively correlated with the amounts of biomass (e. g. $H' / \bar{B} : r = -0,42$, $\alpha < 0,1 \%$). This observation corresponds to the results of GHILAROV & TIMONIN (1972) and the widespread postulate that a system low in diversity needs a higher production rate for its maintainance because of a lower energy utilization (lack of specialists!).

One single sample series (15. 6. 1972, all sand) was used as an example for the comparison between numerical and biomass diversity (Fig. 11). Both indices show the same trend, but it is obvious that biomass diversity ($\bar{H}'_b$ etc.) reaches its constant value much nearer to the outlet due to a high evenness ($\bar{J}'_b$). This early stabilization of the curve is caused by mass populations of very small animals *(Pygospio elegans, Capitella capitata)*. They neutralize the influence of rarer but heavier species on the evenness component of diversity. Except for MCARTHUR's index there is a close relation ($\alpha < 5$ or 1%) between numerically and biomass based diversity values. Biomass diversity seems to be a worse measure for pollution gradients but surely it is a better expression for metabolic processes in the benthic subsystem.

The eel grass community

The eel grass subsystem can only be understood if considered together with the surrounding and adjoining sand bottom. This is altered by the plants in a characteristic way: (1) Roots and a water movement decreased by leaves and blades of *Zostera marina* reduce overturn and transport and firm the bottom. (2) Dead plant parts enrich the sediment with organic substances. (3) The plant itself enables sessile organisms and species preferring phytal and hard bottom to settle on it.

At the same time the typical sand dwelling animals are still present. Eel grass therefore attracts many species from other communities, but contains hardly any specific forms (Table 5). As in the sand bottom community, euryoeceous species dominate the more specific inhabitants (Table 6).

The composition of the feeding types is changed compared with the pure sand community. Most of the species (41 %), but in accordance with ELTON's pyramid only a minority of the mean total individual number (3 %), live as predators. Deposit feeders are still in the forefront (31 % of the species, 67 % of all individuals), but now suspension feeders also play a considerable role (species: 28 %; individuals: 30 %). The rise of predators corresponds to a rise of motile species (63 % of the species number, 20 % of all individuals). Since a *Zostera* subsystem is somewhat stabilized (see above) and contains more (and rarer) species, pollution effects must be assumed to concern not

Table 5

Eelgrass community: Species grouped according to their substrate specifity
(explanations see Table 3)

| a	b	c	d
Lacuna divaricata	*Fabricia sabella*	*Cardium lamarcki*	*Arenicola marina*
	Crangon crangon	*Anaitides mucosa*	*Mytilus edulis*
	Calliopius laeviusculus	*Corophium insidiosum*	*Capitella capitata*
	Rissoa inconspicua	*Asterias rubens*	*Nereis diversicolor*
	Scoloplos armiger	*Gammarus oceanicus*	
	Mya arenaria	*Idotea baltica*	
		Polydora ciliata	
		Microdeutopus gryllotalpa	
		Polydora ligni	
		Littorina littorea	
		Gammarus salinus	
		Hydrobia ulvae	
		Macoma baltica	
		Pygospio elegans	
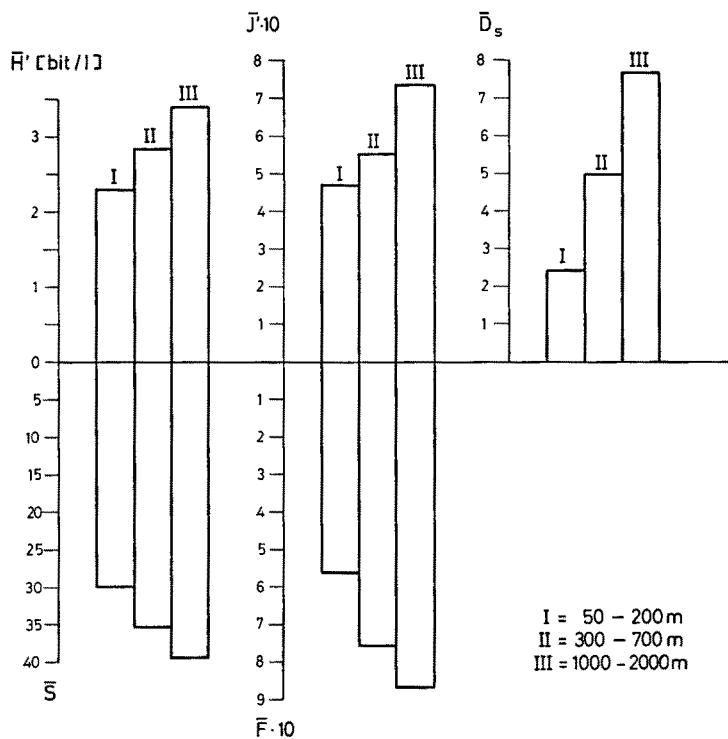
Fig. 12: Eel grass macrofauna: mean number of species $(\overline{S})$ and diversity in the pollution gradient

Table 6

Eelgrass community: rank of species according to their $\overline{A}P^{1/2}$-values (explanations see text)

Species	$\overline{A}P^{1/2}$	$\overline{A}$[l/m²]	$\overline{A}$[l/m²] 50–300 m	A[l/m²] max	$\overline{B}$[mg/m²]	B[mg/m²] max	$\overline{D}$[%]	D[%] max	P[%]
Pygospio elegans	10724	10724	35096	73500	1171,1	12789,0	42,5	88,7	100
Mya arenaria	1699	1806	5194	11680	8483,7	52106,8	5,8	43,3	88,5
Mytilus edulis	1535	1565	4054	14770	19130,0	109591,0	5,7	31,2	96,2
Hydrobia ulvae	1473	1502	3750	9065	798,6	5825,4	6,0	28,6	96,2
Fabricia sabella	1171	1489	0	21110	ca. 11,7	165,5	5,0	50,7	61,9
Oligochaeta	963	963	6768	3080	151,4	737,0	12,6	46,5	100
Capitella capitata	813	833	17623	14500	211,8	2914,5	1,6	16,2	95,2
Polydora ciliata	618	672	14	7300	212,7	2275,0	2,7	22,5	84,6
Scoloplos armiger	374	404	4	1600	845,7	3136,5	3,2	17,5	85,7
Corophium insidiosum	362	362	1042	4100	16,0	61,0	5,3	15,2	100
Macoma baltica	330	330	463	1345	5468,9	29510,5	2,4	10,6	100
Idotea baltica	206	210	440	2380	315,3	1925,0	1,9	15,2	96,2
Microdeutopus gryllotalpa	160	170	101	880	23,4	97,0	1,7	9,3	88,5
Cardium lamarcki	120	129	374	2005	104,9	1641,1	1,2	6,7	87,0
Eteone longa	78	78	24	290	58,3	134,5	1,1	3,2	100
Nemertini	59	64	2	250	25,7	102,7	0,9	7,6	85,7
Nereis diversicolor	56	68	915	850	254,7	3455,0	0,5	2,1	66,7
Gammarus oceanicus	45	56	156	1110	24,1	322,0	0,2	1,5	65,4
Anaitides mucosa	40	41	27	180	44,4	224,5	0,6	5,5	95,2
Asterias rubens	28	31	13	150	1115,7	9852,0	0,6	7,4	82,6
Littorina littorea	28	33	22	125	5939,5	27241,0	0,4	3,0	69,6
Lacuna divaricata	27	34	38	160	49,2	169,5	0,3	2,2	61,5
Arenicola marina	23	29	1	330	40,4	415,0	0,2	0,9	61,9
Gammarus salinus	17	22	282	120	20,8	197,0	0,1	0,9	57,1
Polydora ligni	15	20	464	135	8,1	44,0	0,5	4,4	57,1
Rissoa membranacea	10	15	25	85	9,1	51,0	0,2	1,9	46,2
Rissoa inconspicua	9	15	21	110	2,8	25,0	0,1	0,5	38,5
Calliopius laeviusculus	9	13	19	65	2,9	22,5	0,1	0,7	50,0
Crangon crangon	8	10	4	30	48,0	538,5	0,1	0,6	66,7
Retusa truncatula	6	8	1	85	1,6	17,5	0,1	0,8	47,6
Nephtys caeca	3	5	0	5	80,0	416,0	0,1	0,4	42,9
Gammarellus homari	2	4	0	30	0,9	7,0	0,1	0,5	28,6

only the evenness, but also the species richness component. This supposition is confirmed by the measurement of diversity (Fig. 12). The evennes is lowest in the semi-polluted zone (II) because of mass populations of *Pygospio elegans* occurring here. The mean number of species is more than twice as high as in sand communities. The increase of biomass production by organic enrichment is also notable (Fig. 13).
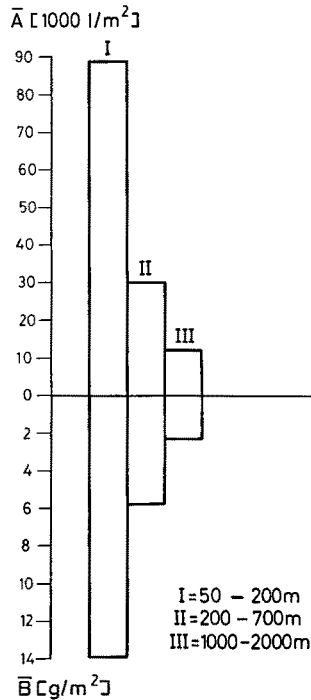

Fig. 13: Eel grass macrofauna: mean abundance $(\overline{A})$ and biomass $(\overline{B})$ in the pollution gradient

Table 6 shows that this can be attributed to the same species as on the sand bottom. The importance for the transformation of organic pollutants and thus for the self purification of the system is of the same order of magnitude.

The mussel bed community

Even more than *Zostera* plants, accumulations of mussels change the sandy biotope: (1) The stabilization of the bottom is highest under a nearly closed cover of shells. (2) Organic substances are enriched in the spaces between the shells. (3) The shells as hard bottom fragments make the formation of phytal possible.

The mussel bed subsystem thus has some peculiarities which allow it to be better described by means of characteristic species. As can be seen in Figures 6 and 7, its macrofauna is much more similar to that of the *Zostera* than that of the sand com-

munity. The main properties of the subsystem are biologically controlled. Table 7 indicates a high amount of characteristic species. Group "a" can also include some rarer species, which also clearly prefer mussel beds as habitat: *Nereis pelagica, Nereimyra punctata, Harmothoe imbricata, Halacarus basteri, Caprella linearis, Brachystomia rissoides.*

Table 7

Mussel bed community: Species grouped according to their substrate specifity
(explanations see Table 3)

a	b	c	d
Balanus crenatus	Polydora ligni	Rissoa inconspicua	Mya arenaria*
Harmothoe impar	Microdeutopus	Fabricia sabella	Nereis diversicolor
Nereis succinea	gryllotalpa	Calliopius	Macoma baltica*
Jaera albifrons	Gammarus salinus	laeviusculus	Pygospio elegans
Pholoe minuta	Corophium	Anaitides mucosa	Hydrobia ulvae
Carcinus maenas	insidiosum	Cardium lamarcki*	Scoloplos armiger
	Asterias rubens	Capitella capitata	
	Polydora ciliata		
	Littorina littorea		
	Gammarus oceanicus		
	Idotea baltica		
	Mytilus edulis		
	Arenicola marina*		

* mostly as juveniles in the habitat

It can be noted that *Mytilus edulis* is the most important but not the most characteristic species (Tables 7, 8). Opportunists also dominate the $\overline{AP}^{1/2}$ rank in mussel beds (Table 8).

The number of feeding types is increased by the parasitic snail *Brachystomia rissoides,* which is, however, rather unimportant. As in the *Zostera* community, predators take the first rank among the species (45 %), but provide only 6 % of the individuals. Deposit feeders (species: 23 %, individuals: 45 %) are displaced from their prominent rank by suspension feeders (species: 29 %, individuals: 54 %). The community structure resembles hard bottom biocoenoses, where the importance of the pelagic system and an unusual richness of species and life forms are also typical.

Sessile filter feeding epifauna *(Mytilus edulis, Balanus crenatus)* is of great importance compared with sediment dominated systems. Motile animals (mostly predators) provide 58 % of all species and 19 % of all individuals.

The percentage of hemisessile infauna in the individuals is decreased from 80 % in pure sand and 73 % in eel grass to only 51 % in mussel beds.

In the pollution gradient an interesting phenomenon becomes obvious: The rise of mean biomass and abundance is much more weakly pronounced here than on the other substrate types (Fig. 14). From Table 8 it can be concluded that the euryoeceous species are not able to increase their population densities to the same extend they do in sandy biotopes. This holds true even for species which prefer mussel beds as habitat (e. g. *Capitella capitata*). The causes and significance of this will be discussed below.
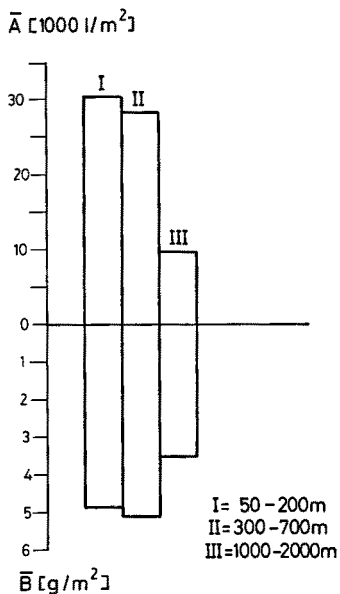
$\overline{A}$ [1000 I/m²]

I= 50 – 200m
II = 300 – 700m
III = 1000 – 2000m

$\overline{B}$ [g/m²]

Fig. 14: Mussel bed macrofauna: mean abundance ($\overline{A}$) and biomass ($\overline{B}$) in the pollution gradient

$\overline{H}'$ [bit/I] $\overline{J}'\cdot 10$ $\overline{D}_s$

I = 50 – 200 m
II = 300 – 700 m
III = 1000 – 2000 m

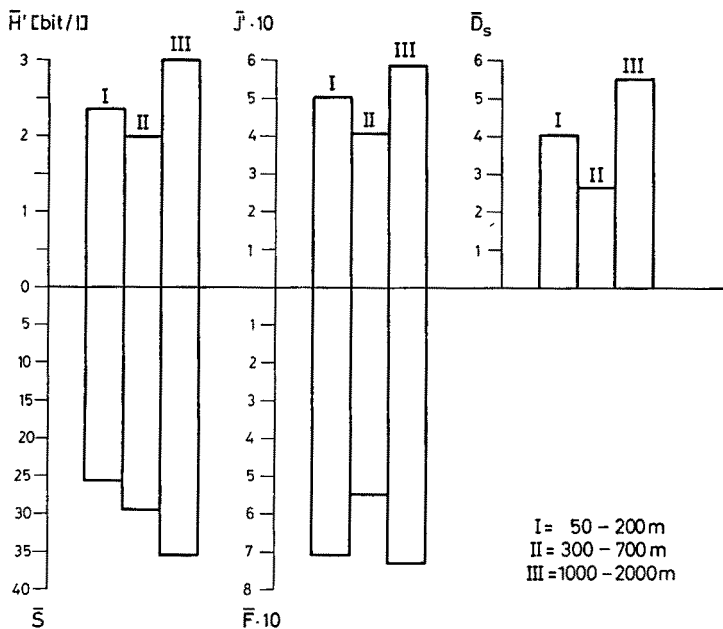$\overline{S}$ $\overline{F}\cdot 10$

Fig. 15: Mussel bed macrofauna: mean number of species ($\overline{S}$) and diversity in the pollution gradient

Table 8

Mussel bed community: rank of species according to their $\overline{A}P^{1/2}$-values (explanations see text)

Species	$\overline{A}P^{1/2}$	$\overline{A}[I/m^2]$	$\overline{A}[I/m^2]$ 50–300 m	$A[I/m^2]$ max	$\overline{B}[mg/m^2]$	$B[mg/m^2]$ max	$\overline{D}[\%]$	$D[\%]$ max	$P[\%]$
Mytilus edulis	5701	5701	7816	22940	965200,0	4143209,0	29,0	83,2	100
Pygospio elegans	5001	5001	21180	71085	287,4	348,2	8,3	54,5	100
Capitella capitata	1830	1830	2859	16637	213,3	1590,0	9,8	35,6	100
Oligochaeta	1767	1767	2643	10780	469,9	3580,0	14,1	72,6	100
Polydora ciliata	1411	1491	267	10420	365,8	2715,0	5,5	36,9	89,5
Corophium insidiosum	1012	1067	4545	2870	93,8	200,0	9,4	21,1	90,0
Microdeutopus gryllotalpa	706	746	686	1900	105,4	289,0	4,7	18,0	89,5
Hydrobia ulvae	574	607	1900	6770	540,9	3493,3	2,4	15,6	89,5
Mya arenaria	374	384	860	3210	3737,8	22886,2	2,8	36,1	94,7
Polydora ligni	294	355	297	1055	189,3	699,0	1,7	6,1	68,8
Gammarus salinus	239	255	401	1740	425,8	2889,0	2,7	15,2	87,5
Fabricia sabella	233	310	966	4825	ca. 6,5	101,3	0,3	3,7	56,3
Idotea baltica	228	234	366	1135	290,4	1277,0	1,1	5,0	94,7
Scoloplos armiger	132	176	282	1200	312,1	2328,7	1,8	14,5	56,3
Macoma baltica	125	125	135	390	3811,6	20330,0	0,8	4,6	100
Nemertini	124	133	186	680	171,8	427,7	1,2	8,2	87,5
Nereis succinea	110	124	37	830	255,7	1536,0	0,7	4,2	78,9
Balanus crenatus	101	113	203	725	2383,0	17514,0	1,3	13,5	78,9
Harmothoe impar	96	96	26	410	61,0	303,0	0,6	1,5	100
Gammarus oceanicus	89	104	184	805	280,9	1732,0	0,8	6,9	73,7
Asterias rubens	68	75	14	310	52403,3	443960,0	0,6	2,8	83,3
Littorina littorea	64	72	63	290	28603,3	135240,0	0,5	3,7	77,8
Nereis diversicolor	54	60	126	245	581,5	2186,0	0,6	2,6	81,3
Jaera albifrons	38	45	64	245	7,9	26,0	0,2	1,4	72,2
Halacarus basteri	36	41	47	123	—	—	0,3	2,1	75,0
Arenicola marina	29	39	33	215	9,8	58,0	0,4	1,9	56,3
Pholoe minuta	23	28	20	100	5,3	27,0	0,2	0,8	68,8
Eteone longa	18	23	24	93	9,2	31,5	0,4	4,0	63,2
Carcinus maenas	18	21	3	120	5298,3	31170,5	0,2	0,9	75,0
Harmothoe imbricata	14	19	13	55	244,6	922,0	0,1	0,5	56,3
Brachystomia rissoides	13	19	5	110	4,1	24,5	0,1	0,7	44,4
Cardium lamarcki	11	15	11	135	26,2	155,0	0,1	1,0	55,6
Anaitides mucosa	9	11	27	80	6,3	55,5	0,1	0,4	62,5
Rissoa inconspicua	8	12	33	60	3,5	18,0	0,1	0,5	42,1
Nereis pelagica	7	9	2	27	81,8	255,5	0,1	0,2	57,9
Calliopius laeviusculus	4	6	18	60	2,3	13,0	0,1	0,7	42,1
Caprella linearis	4	6	13	60	—	—	0,1	0,5	37,5
Nereimyra punctata	1	3	1	20	2,6	17,0	0,1	0,3	21,1

Diversity and mean number of species are highest in mussel beds (Fig. 15). The negative effect of pollution on both species richness and evenness is evident.

COMPARISON OF THE COMMUNITIES AND DISCUSSION

The investigated area belongs to the "shallow inshore benthos" subsystem within Kiel Bay. It contains continua and well defined communities. The pollution gradient creates a continuum, which is partly superimposed on the second type of subordinate subsystems characterized by the substrate properties. Thus the communities of the different substrate types are more similar to each other under the influence of sedimenting sewage particles. In the strongly affected (physically controlled) zone a group of euryoeceous species is predominant, nearly independently of the biotope structure. At greater distances from the sewage outfall the communities are rather more biologically controlled. Mussel beds and eel grass promote similar changes in the sand bottom, which include stabilization and partial covering of the sediment, enrichment of organic material, provision of hard bottom fragments, etc. The consequence is a much greater similarity between the *Zostera* and *Mytilus* communities than occurs between pure sand and any other substrate. The sand and eel grass communities are of course closer related than sand and mussel beds, since the sediment factor is important to a different degree. In Figure 16 these relationships are quantified as reciprocal JACCARD-indices of faunal similarity. The greater the similarity of community composition the nearer two subsystems appear in the triangle.
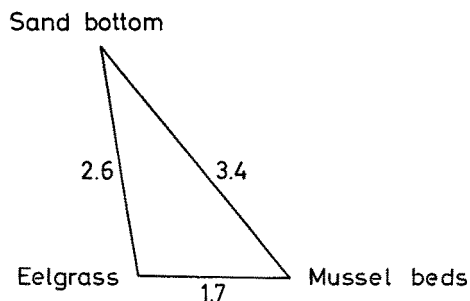

Fig. 16: Faunal similarity of the three main substrate types, expressed as reciprocal Jaccard-index

Comparing the communities as regards life forms in the sequence sand – eel grass – mussel beds the following trends can be observed (Fig. 17, 18): (1) rise of predators, (2) strong decrease of deposit feeders, (3) strong rise of suspension feeders, (4) rise of sessile epifauna, (5) decrease of hemisessile infauna, and (6) rise of motile species (mainly epifauna).

These trends are caused by spatial structure and linked factors. The vertical expansion of the biotope linked to the covering of the sediment increases the importance

of the near bottom water for the benthic fauna. Some typical species are good swimmers (amphipods, isopods, some polychaetes).
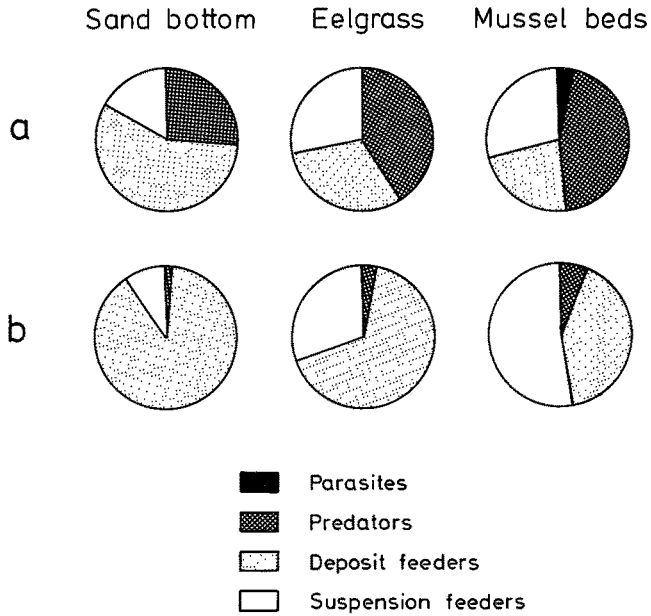

Fig. 17: Substrate types: mean percentage of the main feeding types (a = species, b = individuals)
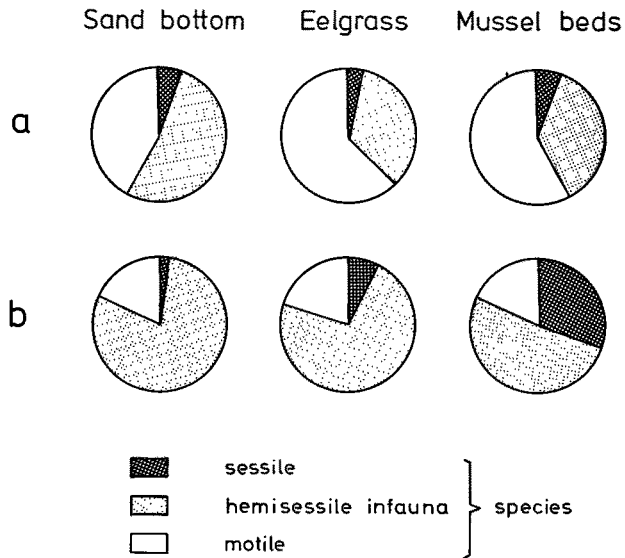

Fig. 18: Substrate types: mean percentage of the settling types (a = species, b = individuals)

The complication and stabilization of the abiotic environment provides a greater capability of accommodating more specialized species in more ecological niches (Fig. 19). The consequence is an increased diversity (Fig. 20).
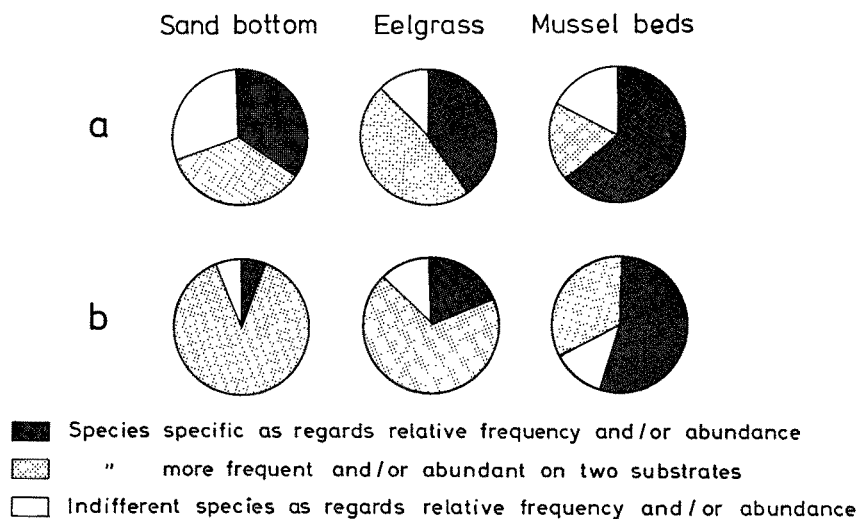
Sand bottom   Eelgrass   Mussel beds

a

b

■ Species specific as regards relative frequency and / or abundance

▨   "   more frequent and / or abundant on two substrates

☐ Indifferent species as regards relative frequency and / or abundance

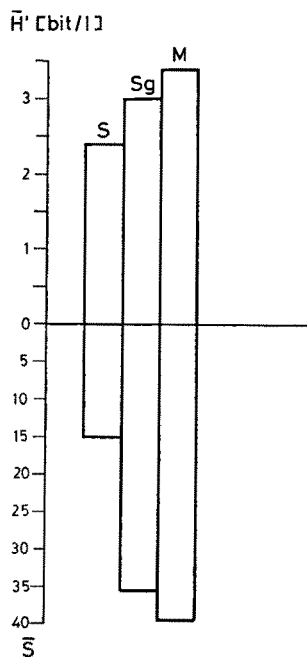Fig. 19: Substrate types: Specifity of the fauna (a = species, b = individuals)

H̄' [bit/I]

Fig. 20: Substrate types: mean number of species (S̄) and diversity

As shown above, sediment communities rather low in diversity react to sedimentation of organic material by producing mass populations of a few generalists. In more highly diverse systems population increases are far weaker. Here strong competition and predation by many specialized species takes place and hinders the development of mass populations. This biological control can be regarded as an ecological buffer of the system against one type of environmental fluctuation: the regulation sets in if those stress factors which selectively favour a few preadapted generalists are rising. Sedimentation of organic debris, for example, favours euryoeceous deposit feeders in particular. Simply organized systems (e. g. sand) are not buffered to this extent. Since they contain mainly generalists, the reactions of the community appear as extreme oscillations of evenness and production.

The trend shown by this comparison confirms partly the diversity – stability hypothesis (McArthur, 1955). Nevertheless it cannot be generalized that higher diversity is basically identical with better self regulation or stronger resistence. Rather the character of the disturbance is decisive: some evidence is given that the relationship stated above is valid only for factors influencing the evenness component of diversity directly (see above), but not for those primarily affecting the species richness component (e. g. poisonous substances). The last mentioned factor type probably has much more catastrophic effects in complex systems containing more specialized and sensitive species than in simpler ones with their mainly euryoeceous species.

These statements upon the relationships among biotope and community structure, environmental stability, diversity and resistance against external disturbances are to be regarded as widely hypothetical rules. They must be scrutinized again in practice to become important for predictions on the behaviour of ecological systems under known conditions.

Bacteria and microfauna, but to a certain extent also, the inshore benthic communities, play an important role for the transformation of particulate organic matter to utilizable biomass (mainly food for flatfishes). Subsystems naturally low in diversity have a far greater capacity than highly diverse ones. The attainable maximum limited by competition for space was observed in sand bottoms as a mean rise of population density by a factor of 50 and of biomass by a factor of 15. A heavier sedimentation of sewage particles decreases the capability of the benthic community to assimilate them (cf. Figs. 1 and 9). The transformation of organic pollutants to available food prevents a suffocation of the marine ecosystem and raises its productivity, but at the same time it creates the danger of accumulation of heavy metals, pesticides and other poisonous substances by the same factor. A second danger, especially for the Baltic Sea, is eutrophication by dissolved nutrients and as a consequence the contamination of the deeper areas by the release of hydrogen sulphide.

From the greater fluctuations of evenness and production caused by disturbances in subsystems of rather simple structure (e. g. sand), it can be deduced that these communities are good indicators for stress factors which affect evenness. Communities of highly complex systems (e. g. mussel beds) are probably better indicators for factors which decrease species richness.

The practical value of this knowledge is that it points out a possible ecological method of measuring environmental stress. It complements physical and chemical

Table 9

Survey of significant species of macrofauna in the study area

Mollusca
Lacuna divaricata FABRICIUS
Littorina littorea L.
Hydrobia ulvae PENNANT
Rissoa membranacea ADAMS
R. inconspicua ALDER
Brachystomia rissoides HANLEY
Retusa truncatula BRUGIERE
Mytilus edulis L.
Cardium lamarcki REEVE
Macoma baltica L.
Mya arenaria L.

Annelida
Harmothoe imbricata L.
H. impar JOHNSTON
Pholoe minuta FABRICIUS
Eteone longa FABRICIUS
Anaitides mucosa OERSTED
Nereimyra punctata O. F. MÜLLER
Nereis pelagica L.
N. diversicolor O. F. MÜLLER
N. succinea LEUCKART
Nephtys caeca FABRICIUS
Scoloplos armiger O. F. MÜLLER
Spio filicornis O. F. MÜLLER
Polydora ligni WEBSTER

P. ciliata JOHNSTON
Pygospio elegans CLAPAREDE
Capitella capitata FABRICIUS
Arenicola marina L.
Fabricia sabella EHRENBERG

Acari
Halacarus basteri JOHNSTON

Crustacea
Balanus crenatus BRUGIERE
Gammarus oceanicus SEGERSTRÅLE
G. salinus SPOONER
Calliopius laeviusculus KRÖYER
Bathyporeia pilosa LINDSTRÖM
B. sarsi WATKIN
Microdeutopus gryllotalpa COSTA
Corophium insidiosum CRAWFORD
Caprella linearis L.
Pseudocuma longicorne BATHE
Idotea baltica PALLAS
Jaera albifrons LEACH
Crangon crangon L.
Carcinus maenas L.

Echinodermata
Asterias rubens

evaluations, since the community structure reflects the long term situation of its biotope by effectively carrying out the integration of all factors which otherwise must be continuously measured over a long period and fed into a computer. The precondition for using benthic communities as indicators of environmental quality is the further study of all basic structural and functional relationships. The last step will be the building of a comprehensive model for the description, understanding and simulation of the endangered ecosystem.

SUMMARY

1. The influence of organic sewage pollution on the macrobenthic community structure of a shallow inshore region was studied in the South of Kiel Bay (Baltic Sea).
2. The sandy bottom bears three associations within the pollution gradient. They can be characterized by: (a) *Capitella capitata* and Oligochaeta (50–100 m distance from the sewage outlet), (b) *Pygospio elegans* (200–ca. 700 m) and (c) *Bathyporeia sarsi* (> 700 m). They have no distinct borders but resemble rather a continuum.
3. In the less affected reaches the influence of biological substrate structure becomes more evident. Distinct associations can be distinguished.

4. The combination of life forms, number of species, diversity and other community features depend on the degree of the substrate's spatial heterogenity in a characteristic way. They are also strongly influenced by the extent of pollution.

5. The inshore benthic macrofauna provides an important contribution to the self purification of the whole ecosystem by transforming particulate organic matter to available fish food. Abundance and biomass are multiplied due to sewage sedimentation. This process creates the danger of accumulation of poisonous substances in addition to the eutrophication problem.

6. More highly diverse systems (e. g. mussel beds) resist better those factors which directly influence the evenness of the community. The ecological buffer consists of many specialized competitors and predators. Systems low in diversity (e. g. sand) react much more strongly to disturbances of this type.

7. Faunal composition, evenness, species richness, population density and biomass of the communities are suitable parameters of an ecological method for the early recognition of environmental stress.

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