

## On the population development of the introduced razor clam *Ensis americanus* near the island of Sylt (North Sea)

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**ABSTRACT:** The American razor clam *Ensis americanus* (= *E. directus*) was introduced into the eastern North Sea in the late 1970s. By larval and postlarval drifting the species rapidly extended its distribution, now ranging from the English Channel to the Kattegat. Near the island of Sylt in the eastern North Sea it has been recorded since 1979. Recruitment was rather irregular, with about six strong year-classes within two decades. Growth seems comparable with populations in its native range (Atlantic North America). Although present in the lower intertidal zone, maximum densities occurred in shallow subtidal sand with a biomass similar to that of dense beds of native cockles and mussels in the adjacent intertidal zone. *Ensis americanus* established in otherwise sparsely faunated sand (channels exposed to strong currents) as well as in dense infaunal assemblages (lower intertidal and subtidally). There were no significant interactions with resident species. In dense beds of razor clams, however, fine sediment particles accumulated which may have altered abundances of polychaetes. In spite of high annual variability, *E. americanus* has become a prominent component of the macrobenthos in shallow subtidal sands of the North Sea.

### INTRODUCTION

The American razor clam *Ensis directus* (Conrad, 1843) is a well-known inhabitant of the lower intertidal and shallow subtidal zones along the entire US Atlantic coast between Labrador and Florida (Theroux & Wigley, 1983). According to Van Urk (1964, 1987) the name *Ensis directus* refers to a Miocene fossil while the recent species should be properly named *Ensis americanus* (Binney, 1870). We prefer to use the latter name.

*Ensis americanus* (hereafter abbreviated as *Ensis*) was introduced to Europe in the late 1970s, presumably as larvae in the ballast water of a ship crossing the Atlantic (Von Cosel et al., 1982). Due to a long-lasting pelagic occurrence of larvae and byssus-drifting post-larvae, the species rapidly spread over the continental coastline of the North Sea from Denmark to northern France (Essink, 1985, 1986; Luczak et al., 1993). Recently it reached the British Isles and the west coast of Sweden (Howlett, 1990; Jansson, 1994). In the Netherlands, highest abundances were recorded in lower intertidal and shallow subtidal areas with a rather mobile sediment (Beukema & Dekker, 1995). These areas were scarcely occupied by native fauna, while *Ensis* managed to overcome frequent sediment disturbance by burying both fast and deep (Swennen et al., 1985;

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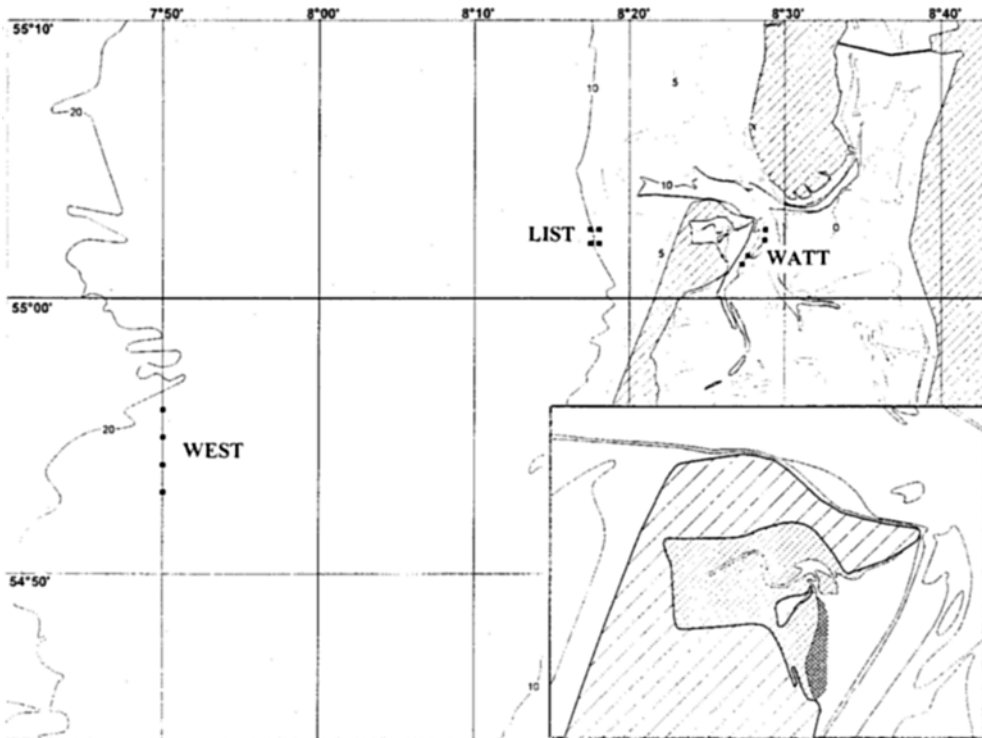


Fig. 1. Localities (West, List, Watt) near the island of Sylt regularly sampled for *Ensis americanus*. Inset shows the northernmost part of Sylt (hatched) with the Königshafen intertidal flats (shaded) and studied area just below spring low tide level (crosswise hatched)

Schiedek & Zebe, 1987). Using records from the coastal waters around the island of Sylt we studied the development of the newcomer in the northern Wadden Sea, including coastal fine-sand habitats rich in other macrobenthic species. Will *Ensis* be able to successfully compete in habitats densely populated by other macrofauna and with abundant epibenthic predators?

#### MATERIALS AND METHODS

From 1992 to 1996, the development of the macrobenthic fauna was studied in three areas near the island of Sylt in the North Sea (West, water depth 20 m; List and Watt, water depth 10 m; see Fig. 1). The sediment is composed of fine to medium sand (median diameter of sand grains 150–200  $\mu\text{m}$ ). In each area four sites were sampled, 1–2 km apart (Fig. 1). Using a Reineck box corer, six sediment samples of 0.02 m<sup>2</sup> were collected from each of the sites, in March and September. The sediment was sieved through 1-mm<sup>2</sup> meshes and the residue preserved in 5% buffered formalin/seawater solution. From these samples the macrofauna was identified to the lowest possible taxonomic level, counted, and bivalves measured to the nearest millimetre. In addition to these subtidal sites we studied an area just below spring low tide level east of Sylt (Fig. 1, in-

set). Here, 15 cores of 0.02 m<sup>2</sup> were collected by hand in March 1993, when strong offshore winds exposed this area.

The intertidal distribution of *Ensis* was studied in August/September 1989 (completed by sampling a few plots in August 1990) by mapping the tidal flats of Königshafen (Fig. 1) based on a grid of 100 × 100 m plots ( $n = 450$  plots). Every second plot (total  $n = 222$ ) was studied by 20 times excavating the sediment from 30 × 30 cm areas with a spade. The sediment was searched by hand and the species visible to the naked eye were recorded in a presence/absence sheet. These 20 samples from each of the plots yielded a semi-quantitative estimate of species abundance (range 0–20 occurrences per plot). Since its first appearance in 1979, qualitative records on the occurrence and size of *Ensis* have been taken all around the island of Sylt.

## RESULTS

### Recruitment

Recruitment showed strong year-to-year variability. Since its first spatfall in 1979, strong cohorts originated in 1981, 1986, 1987, 1990, and 1994 in the lower intertidal zone. In the subtidal Wadden Sea (site Watt), an outstanding recruitment occurred in 1993 when some 1900 *Ensis* m<sup>-2</sup> accumulated until September (see Table 1), 57% of which survived their first winter. With an average of 1500 recruits m<sup>-2</sup> the same year class was successful in the subtidal North Sea (site List, see Fig. 1) as well.

### Abundance and biomass

In the Königshafen intertidal, *Ensis* large enough to be detected with the naked eye were essentially restricted to the flats below mean tidal level (Fig. 2) and abundance increased towards low tide level. The maximum intertidal density of *Ensis* (33 m<sup>-2</sup>) was found just below spring low tide level in March 1993 (crosswise hatched area in Fig. 1). This population was composed of 0.5- to 2.5-year-old specimens and had a biomass of 78 g ash free dry weight (AFDW) m<sup>-2</sup>. This amounts to 66% of the total macrofaunal biomass at this low intertidal site.

In the subtidal, recruit abundances of up to 2000 m<sup>-2</sup> occurred at least once in all three areas (Table 1). Peaks of biomass were associated with cohorts 1.5 or 2.5 years old and reached values as high as 667 g AFDW m<sup>-2</sup> in the Wadden Sea and 250 g AFDW m<sup>-2</sup> in the North Sea. No other species reached a comparable biomass at the same time; hence *Ensis* was the prominent species (Table 1). Within single sites, peaks of abundance and biomass even attained twice these values (maximum recorded biomass 1400 g AFDW m<sup>-2</sup>, which may be equivalent to some 15 kg FW m<sup>-2</sup>).

### Growth

The size of *Ensis* varied strongly between years and sites (Table 2). Average size of 1-year-old individuals did not significantly differ between the List site west of the island and the subtidal Wadden Sea (Watt). However, 2-year-old specimens grew larger in the Wadden Sea (Table 2). In the intertidal zone, the 1979 cohort stayed unusually small

Table 1. Abundance ( $n\ m^{-2}$ ) and biomass ( $g\ AFDW\ m^{-2}$ ) of *Ensis americanus* in three subtidal areas near the island of Sylt, September 1992–1996. Numbers in parentheses give percentage share of *Ensis* biomass in macrobenthos communities. See Fig. 1 for localities

Site Year	West		List		Watt	
	Abun- dance	Biomass	Abun- dance	Biomass	Abun- dance	Biomass
1992	< 10	< 0.1 (< 1%)	19	0.6 (3%)	< 10	< 0.1 (3%)
1993	< 10	< 0.1 (< 1%)	1500	18.5 (45%)	1925	38.1 (97%)
1994	< 10	< 0.1 (< 1%)	2025	249.9 (93%)	1159	36.8 (99%)
1995	< 10	< 0.1 (< 1%)	204	22.2 (73%)	467	468.8 (99%)
1996	1883	6.2 (22%)	38	13.7 (65%)	688	666.7 (99%)

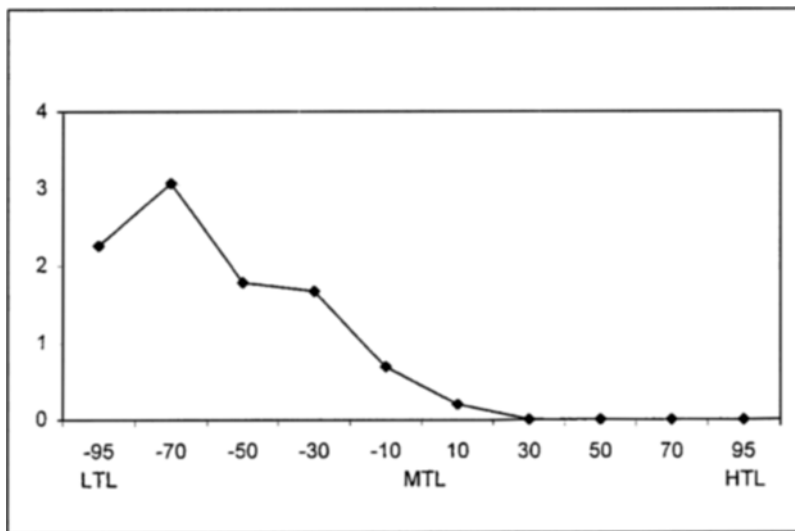


Fig. 2. Mean abundance of *E. americanus* in tidal levels of Königshafen (August/September 1989, August 1990). Abundance was estimated by the mean number of occurrences per studied plot (presence, range 0–20), here averaged over the plots within each height interval. The total number of studied plots was 222. LTL, low tide level; HTL, high tide level; MTL, mean tidal level

during their first year; possibly these specimens derived from a late spatfall. During their second year, however, they achieved average size. The largest individual found had a shell length of 186 mm and an estimated age of 7 years. However, most cohorts disappeared as they attained an age of 2–4 years.

### Mass mortality

In February 1991 and again in March 1994 we observed masses of mostly 4-year-old *Ensis* protruding above the sediment by about half their length in the intertidal zone (Fig. 3). Some of them rapidly re-buried when we approached, but most of them showed no reaction, even after touching. During the following weeks, these organisms died and

Table 2. Average shell length (mm) of cohorts of *Ensis americanus* in the North Sea

Site	Age				Reference
	1 year	2 years	3 years	4 years	
DK, Blåvandshuk beach	48	109	137	152	Mühlenhardt-Siegel et al., 1983
D, Sylt	43	94	124		Mühlenhardt-Siegel et al., 1983
D, Sylt intertidal: cohort 1979	21	92			This study
D, Sylt intertidal: cohort 1990	43	97	127	133	This study
D, Sylt WATT	34	82– 92			This study
D, Sylt LIST****	26–39	67– 81			This study
D, Sylt WEST	15*				This study
D, Langeness intertidal**	27–30	78– 84			Swennen et al., 1985
D, Großer Vogelsand***	10–50	60–105			Dörjes 1992
NL, lower intertidal	64	126	144		Beukema & Dekker 1995
F, Dunkerque	35*				Luczak et al., 1993

DK, Denmark; D, Germany; NL, the Netherlands; F, France  
 \* age 0.5 years  
 \*\* range of averages from three sites  
 \*\*\* range of averages from three sites studied over 10 years, estimated from graph  
 \*\*\*\* range of averages from 4 years

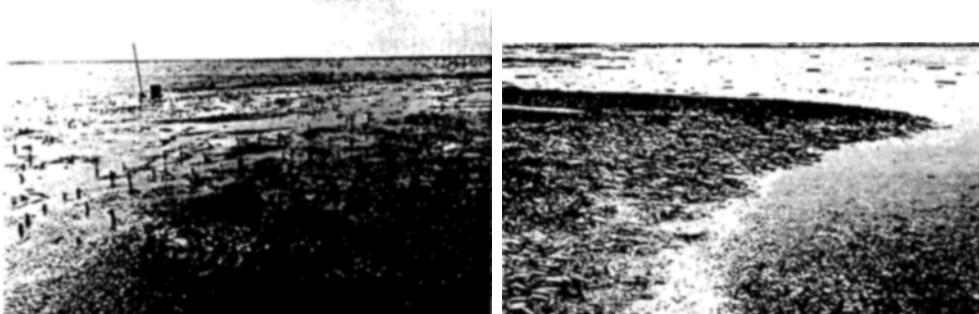


Fig. 3. Four-year old *E. americanus* protruding above the sediment after a cold spell in March 1994 (left) and later washed shoreward, here accumulating at a mussel bed (right) in Königshafen, island of Sylt

got washed ashore. Mass mortality also occurred in cohorts of < 1-year-old specimens, i.e. in the lower intertidal zone of Königshafen in March 1980 and February 1982.

## DISCUSSION

### Recruitment

*Ensis* release eggs in March and April (Mühlenhardt-Siegel et al., 1983) and spatfall occurs after a planktonic phase of 2–4 weeks. However, in the northern Wadden Sea there are several periods of spatfall (Armonies, 1992, 1996). This may have caused the several

peaks in the size distribution of single year-classes observed by Mühlenhardt-Siegel et al. (1983). In the Sylt data the first peak was always the highest (Armonies, 1992, 1996). Unpublished data on other bivalves in the Sylt area indicate that this may be a rather general pattern of reproduction.

Near the island of Sylt initial spatfall mainly occurred in the lowermost intertidal and shallow subtidal zone while byssus-drifting post-larvae moved towards deeper waters during the following weeks (Armonies, 1996). During these migrations many juveniles intermittently showed up in upper intertidal flats but did not stay there. In laboratory experiments they avoided the burrowing and defaecation activity of lugworms *Arenicola marina* (Dölle, 1996). Thus, less recruitment in higher than in lower parts of the intertidal zone as observed by Beukema & Dekker (1995) is not due to a lack of potential settlers but indicates habitat selection by drifting juveniles (Armonies, 1998).

Recruitment generally showed a strong year-to-year variability (Table 1). In the Dutch intertidal zone, a strong recruitment occurred only once in 13 years (Beukema & Dekker, 1995). Of these recruits only 4% survived their first winter. Generally, over-winter survival decreased with increasing tidal elevation. However, since juveniles up to 55 mm in length have been observed drifting in the water column (Essink, 1985, and references therein) the specimens might in part have left the upper intertidal sites, resulting in skewed estimates of local survival. Similar winter migrations also occur in other bivalves like *Macoma balthica*, leading these organisms from the high intertidal towards lower tidal elevations (Beukema, 1993).

A low over-winter survival (1-7%) was also reported for subtidal populations in most of the years, while Mühlenhardt-Siegel et al. (1983) reported on a single year with a survival as high as 55%. The situation is similar in the Sylt area where a high recruitment and a high over-winter survival were only observed in 1993/1994 (Watt and List sites). However, the burrowing capacity of *Ensis* might strongly skew abundance estimates as well as derived parameters like mortality. Broken individuals in our samples indicate that penetration depth of our box corer was insufficient for the vertical burrowing capabilities even of only 1-year-old *Ensis*. As larger individuals are even deeper burrowers (Swennen et al., 1985) abundance of older year classes may mostly be underestimated. In addition, because of seasonal differences in bioturbation activity and spatial differences in sediment composition, average penetration depth of our box corer varied with the season and the sites. This results in corresponding differences in sampling efficiency of the deep-dwelling fauna.

Both a high recruitment in few years only and a low average winter survival are rather common among Wadden Sea bivalves (Beukema, 1990, and references therein). However, a high mortality, particularly of the youngest size class, by crab predation in late summer is also quite common in the Wadden Sea. Therefore the timing of sample collection is crucial for discriminating between late summer mortality and over-winter survival in a strict sense.

### Abundance and biomass

In intertidal populations abundance was found to correlate negatively with increasing tidal height. Beukema & Dekker (1995) reported a maximum density of 144 m<sup>-2</sup> (recruits) and a maximum biomass of 17.4 g AFDW m<sup>-2</sup> (in a cohort of 2.5-year-old speci-

mens). The maximum density found in the Sylt intertidal zone was lower, but since the Sylt population was composed of specimens up to 3.5 years old, biomass was higher (78 g AFDW m<sup>-2</sup>).

Subtidal populations seem to prefer a water depth < 18 m (see below) while the maximum recorded locality in the North Sea was as deep as 26 m (Mühlenhardt-Siegel et al., 1983). Shortly after spatfall these authors found an abundance of 440 recruits m<sup>-2</sup> in the entire area studied and a local maximum of 15 440 m<sup>-2</sup>. These figures surpass the respective intertidal values by an order of magnitude. Due to low survival, however, biomass (24 g AFDW m<sup>-2</sup>) did not exceed intertidal values. In the subtidal near the island of Sylt, recruit abundances of some 2000 m<sup>-2</sup> occurred in all three areas during one or several years. Peaks of biomass were associated with cohorts 1.5 or 2.5 years old and reached values as high as 667 g AFDW m<sup>-2</sup> (Table 1). These values are similar to cockle (*Cerastoderma edule*) or mussel (*Mytilus edulis*) beds in the same area (Reise et al., 1994; Reise & Lackschewitz, 1998).

### Growth

In the intertidal zone, growth of juvenile *Ensis* correlated negatively with tidal elevation (Beukema & Dekker, 1995). However, this does not mean that growth was still faster in subtidal habitats (see Table 2). Since the lowest growth in the subtidal was found in the West site with an average water depth of 20 m, optimum growth conditions are expected to occur somewhere between 0 and 20 m depth in the North Sea. However, at the List site (depth 10 m) the size of 1-year-old specimens varied strongly between years indicating that other factors than depth and latitude influence growth as well. Presumably, food and temperature will affect growth in *Ensis* in the same strong way as is the case in other bivalves of the Wadden Sea. As *Ensis* occur down to 100-m depth along the US Atlantic coast (Theroux & Wigley, 1983) it should be able to occupy the entire German Bight in the North Sea. However, the water column in this area is frequently density stratified in summer, with a reduced oxygen concentration in the bottom water layer (Hickel et al., 1989). Therefore the 'limited tolerance to hypoxia' noted by Schiedek & Zebe (1987) might restrict *Ensis* to the shallow coastal zones without a thermohaline stratification.

The maximum shell-length so far reported from European populations is 16–17 cm (Beukema & Dekker, 1995). The largest individual found in the Sylt area (18.6 cm, 7 years old) does not greatly increase this range. Apart from exceptions like this, the average age of adult *Ensis* populations in the North Sea seems to be in the range 2–4 years. This conclusion is also supported by shell material found washed ashore.

### Mass mortality

Conspicuous events of mass mortality, mainly in late winter and early spring, accompanied the invasion of *E. americanus* into the North Sea from its very beginning (Mühlenhardt-Siegel et al., 1983). These authors suspected diseases, parasitism, lack of food in high-abundance areas, and energy depletion after spawning as potential causes. Lauckner (personal communication), however, found no parasites or signs of diseases. A lack of food is also not a very convincing hypothesis because in winter en-

ergy demand is usually low while POC (particulate organic carbon) attains its annual maximum. On the other hand, *Ensis* seems to be unaffected by the summer minimum in phytoplankton biomass. Energy depletion after spawning might be a possible cause for mass mortality of older year-classes in early spring but is an unlikely explanation for mortality of specimens < 1 year old.

Two other possible causes may be added here. First, the European *Ensis* population might have derived from an American population at the southern limit of its distributional range which may not be adapted to low winter temperatures. Second, the original American habitats of the invaders are expected to be less affected by winter storms because prevailing westerly winds drive the waves in an offshore direction on the US Atlantic coast but in an onshore direction in the eastern part of the North Sea. As a result, the probability of being washed out of the sediment may be much higher in the European habitats than in the American ones. It is an open question, however, whether *Ensis* retain their fabulous burying capabilities once the ambient temperature approaches the freezing point. Thus we speculate that a combination of frequent sediment disturbance and a generally low activity in winter may be an explanation for the events of mass mortality.

#### Effects on resident communities

While *E. americanus* was mostly found in lower intertidal and shallow subtidal sandy sediments that are relatively exposed to wave action and poor in macrobenthos (Beukema & Dekker, 1995) it also occurred in a species-rich fine-sand *Macoma balthica* – *Tellina fabula* community near Sylt (List site, water depth 10 m). In this area there was only a single significant negative correlation with abundance of another bivalve (within sites, between years), viz. between the abundance of *Ensis* and cockles *Cerastoderma edule*. The latter species had an exceptionally good recruitment in 1996; then, there were almost no recruits in *Ensis* (Fig. 4). Positive correlations between *Ensis* and other infaunal species, on the other hand, were numerous and mostly seem to mirror a common response to winter temperature. Thus, there is no convincing evidence for quantitatively important interactions of the newcomer with the resident fauna and we conclude that *Ensis* was not competitively displaced into mobile sands. In fact, its niche in European waters seems to match the distribution along the US Atlantic coast. Since it thrives southward to Florida and down to a depth of 100 m (Theroux & Wigley, 1983) further expanding of its distributional range in Europe is expected.

Dense aggregates of *Ensis* may take some 5–10% of the superficial sediment volume that is then no longer available to other species. In fact, the mean of abundance of small polychaetes was decreased by some 10% in dense *Ensis* beds compared to neighbouring sites with a low *Ensis* abundance (due to small-scale spatial patchiness, however, this is statistically not significant,  $p > 0.1$ ). Furthermore, adding a high number of large filter feeders to a community will decrease the amount of food for other filter feeders. This may be important in the case of food limitation. Eutrophication during the past decades coincided with increases both in primary and secondary production (Beukema, 1991). This indicates that at least some parts of the coastal zone of the North Sea may indeed be food-limited, though nowadays at a relatively high level. Thus *Ensis* may decrease the stocks of other filter feeders such as cockles and mussels which might be



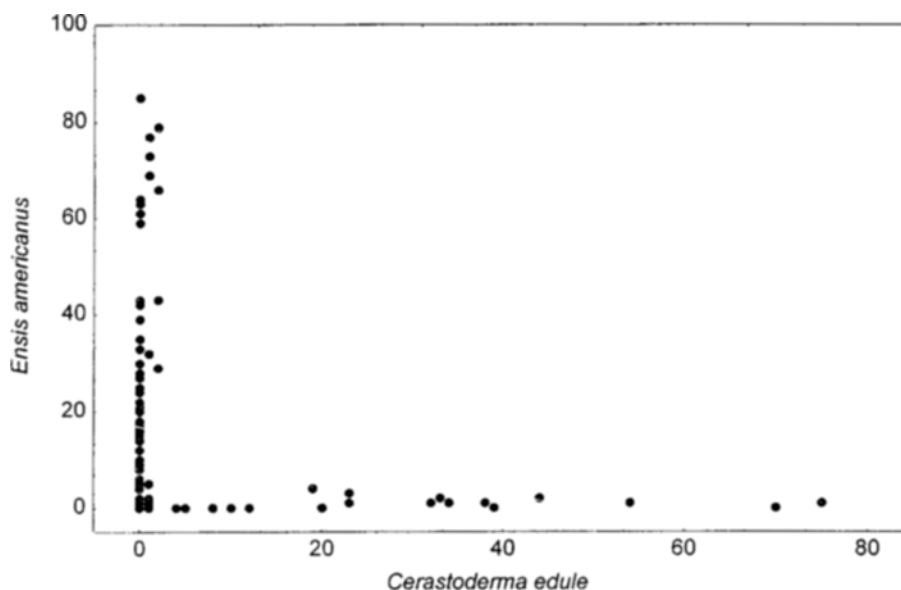


Fig. 4. Correlation between the abundances ( $n$  per box core of  $0.02 \text{ m}^2$ ) of *E. americanus* and *Cerastoderma edule* in subtidal fine sand (List site, water depth 10 m; 24 box cores per year collected in September, 1992–1996. *Cerastoderma edule* abundances  $\geq 5$  all come from year 1996

come particularly important as nutrient levels are managed to decrease to a more natural (pre-industrial) level.

At the Watt site we observed that the sediment containing dense aggregates of *Ensis* tended to become enriched with faecal material. Sediment disturbance and megaripple migration during storm tides removed this material from the top few centimeters of sediment but it was retained in deeper sediment layers. Although this is a mere correlation and it is not known whether this material was released by *Ensis*, within or above the sediment, or only accumulated in *Ensis* beds, changes in the habitat may be expected. Sediment enriched with faecal material will have different mechanical properties, particularly concerning erodibility. This may be why at one site the percentage of very fine sand (particles  $< 0.125 \text{ mm}$ ) increased from an average of 5% in 1992 to some 15% in 1996. During the same period the sediment package became denser, as indicated by decreasing penetration depth of our box-corer. Therefore, the fauna is expected to change as well, in the long run. We have already observed a tendency for change in three polychaetes. *Magelona mirabilis* decreased ( $104 \rightarrow 4 \text{ m}^{-2}$ ) while there were increases in *Capitella capitata* ( $0 \rightarrow 192$ ) and *Capitella minima* ( $0 \rightarrow 102 \text{ m}^{-2}$ ) in these *Ensis* beds. Thus *Ensis* might not only enrich the formerly less populated mobile sands by its own presence, but also indirectly change the associated communities towards a higher diversity.

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