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Macrofauna biodiversity of mussel bed assemblages in Thermaikos Gulf (northern Aegean Sea)

Received: 20 February 2003 / Revised: 22 October 2003 / Accepted: 10 November 2003 / Published online: 12 December 2003
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Abstract Biomonitoring of mussel bed assemblages can provide valuable information about the impact of pollution on hard substrate assemblages. This study of *Mytilus galloprovincialis* mussel beds in Thermaikos Gulf (northern Aegean Sea) deals with the spatial and temporal structure of the associated fauna. Samples were collected and abiotic factors were measured in two successive years. Common biocoenotic methods were employed to analyze the data. The samples could be separated into three groups, with summer and winter samples being clearly different. A total of 100 species were found: polychaetes and crustaceans were the most dominant taxa. The assemblage shows high diversity with respect to species abundance. Biotic interactions within the assemblage appear to influence its composition, although the total evenness remains unaffected in space and time. The *M. galloprovincialis* assemblages can be found in clean as well as in polluted waters and, therefore, are of great interest in biomonitoring studies.

Keywords Infralittoral · Hard substratum · Mussel beds · Biomonitoring

Introduction

Mussel bed (*Mytilus galloprovincialis* LMK) assemblages can develop in clean and moderately polluted as well as polluted waters (e.g. Bellan-Santini et al. 1994; Damian-

idis and Chintiroglou 2000). Therefore, biomonitoring of these assemblages can provide valuable information about the impact of pollution on hard bottom communities (e.g. Wenner 1988).

There is adequate information about the structure of *M. galloprovincialis* assemblages from various regions in the Mediterranean Sea, especially from the western Mediterranean coasts (Bellan 1969, 1980; Bellan-Santini 1969, 1981; Desrosiers et al. 1982, 1986; Hong 1983; Tursi et al. 1984; Tsuchiya and Bellan-Santini 1989), while information on the Aegean Sea is relatively limited (Kocatas 1978; Topaloglou and Kihara 1993; Lantzouni et al. 1998; Damianidis and Chintiroglou 2000). Many authors have put emphasis on the importance of these assemblages in biomonitoring studies (see Thiel and Ullrich 2002). Damianidis and Chintiroglou (2000) reported that the abiotic factors at all sampling sites within Thermaikos Gulf do not fluctuate in time. Therefore, any variation in the composition of *M. galloprovincialis* assemblages has to be attributed to biotic factors.

This study aims at the investigation of the spatial and temporal structure of the fauna associated with *M. galloprovincialis* assemblages on the eastern coast of Thermaikos gulf.

Methods

Study area

The sampling sites were selected on the basis of their historical background, as well as the exposure of the assemblages and the depth of their occurrence. Of all locations with artificial hard substrate along the east coast of Thessaloniki Bay, two appeared to be very similar in bathymetric distribution and exposure of *M. galloprovincialis* populations: the piers of Agia Triada (ST1) and Perea (ST2) (Fig. 1). These piers were constructed 25 years ago and are supported by concrete pillars, which comprise the substrate for the mussel beds. At these sampling sites, the *M. galloprovincialis* populations develop a uniform physiognomic aspect, with large numbers of mussels covering an area that extends from the lower infralittoral zone down to a depth of 2.5 m.

Communicated by H.-D. Franke

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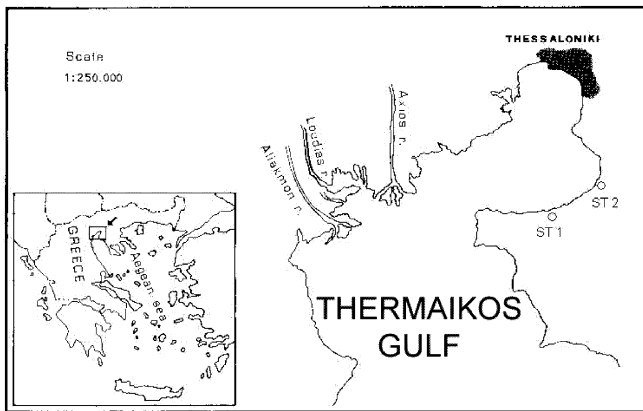


Fig. 1 Map showing Thermaikos Gulf and the two sampling sites (ST1, ST2)

Physico-chemical factors

During this study, physico-chemical factors such as salinity (‰), conductivity ($\mu\text{S}/\text{cm}$), water clarity (m), dissolved oxygen (mg/l), temperature ($^{\circ}\text{C}$) and total hydrodynamics were measured. All measurements were made using WTW (Wissenschaftlich-Technische Werkstaetten, Weilheim, Germany) and Lovibond Checkit (Dortmund, Germany) micro-electronic equipment and water clarity was examined using the Secchi disc. These measurements were conducted monthly at each sampling site and the results have already been reported by Damianidis and Chintiroglou (2000).

Sampling methods

Sampling was carried out while scuba diving. Samples were taken as described by Chintiroglou and Koukouras (1992). The area covered by the quadrat sampler was 400 cm^2 (Stirn 1981; Bakus 1990). Three replicates were taken each time. The samples, 24 in total, were collected during winter and summer of 1994 and 1995. After sampling, the specimens were preserved in 10% formalin and were transferred to the laboratory for further treatment. All samples were collected by the same scuba diver.

Data analysis

Common biocoenotic methods were employed to analyze the faunal composition of the *M. galloprovincialis* assemblages (Bellan-Santini 1981; Damianidis and Chintiroglou 2000). Hence, the numerical abundance (N) on a scale of 1 m^2 , the mean dominance (D) and the frequency (f) were estimated. Also Shannon-Weaver's (H'), and Margalef (d) and Pielou's Evenness (J') were calculated on a \log_2 basis (Daget 1979).

Seasonal differences in the mean numbers of individuals were tested using one-way ANOVA and multiple comparisons. All data were then converted to logarithms. The Spearman rank correlation coefficient (r_s) was employed to determine the relation of the number of mussels (Ab_M) with faunal abundance (mAb) and richness (R).

The numerical abundance data, obtained per sampling station, were analyzed using cluster and multidimensional scaling (mds) techniques, based on the Bray-Curtis similarity, using the PRIMER package (see Clarke and Green 1988; Clarke and Warwick 1994). The square root transformation was applied in order to increase the contribution of the rare species (Clarke and Warwick 1994). The significance of the multivariate results was assessed using the ANOSIM test. SIMPER analysis was performed in order to identify the percentage contribution of each species to the overall similarity within a site and the dissimilarity among sites (Clarke 1993). The

above were carried out to examine the similarity degree of samples, in both space and time.

Results

Composition of the assemblage

A total of 100 species were found associated with *M. galloprovincialis* assemblages in Thermaikos gulf (Table 1). The distribution of these species in major taxonomic groups is given in Table 2. The dominant groups are polychaetes (37.5%) and crustaceans (30.9%). As shown in Table 1, 17 species were distinguished as "very common" ($f \geq 50\%$), 26 as "rare" ($f < 10\%$), and 57 as "common" ($10\% < f < 50\%$).

Among the very common species, the polyclad *Stylochus* sp. and the decapods *Pilumnus hirtellus* and *Pisidia longicornis* have been described as the main predators of mussels (see Galleni et al. 1977; Damianidis and Chintiroglou 2000). The cirripeds *Balanus perforatus* and *B. trigonus*, the polychaete *Serpula vermicularis* and Bryozoa sp1, are well-known as organisms that often settle on mussel shells (Bussani 1983; Damianidis and Chintiroglou 2000). The rest of the very common species are free motile organisms which employ various modes of feeding. The tube-building peracarids are detritivores (Barnard 1958, 1963; Isaac et al. 1994), *Ophiothrix* sp. (Pérès 1976) is a suspensivore, and the polychaete *Staurocephalus rudolphii* is a carnivore (Fauchald and Jumars 1979).

Relationships between fauna and structure of mussel beds

According to Tsuchiya and Nishihira (1986), the morphology and relative age of the mussels in an assemblage can play a significant role in the composition of the associated fauna. Therefore, the correlations of mussel abundance with species richness and faunal abundance, respectively, were examined.

For both summer and winter samples, faunal abundance (mAb) and richness (R) were not correlated ($P > 0.5$) with the mussel abundance (Ab_M).

Diversity

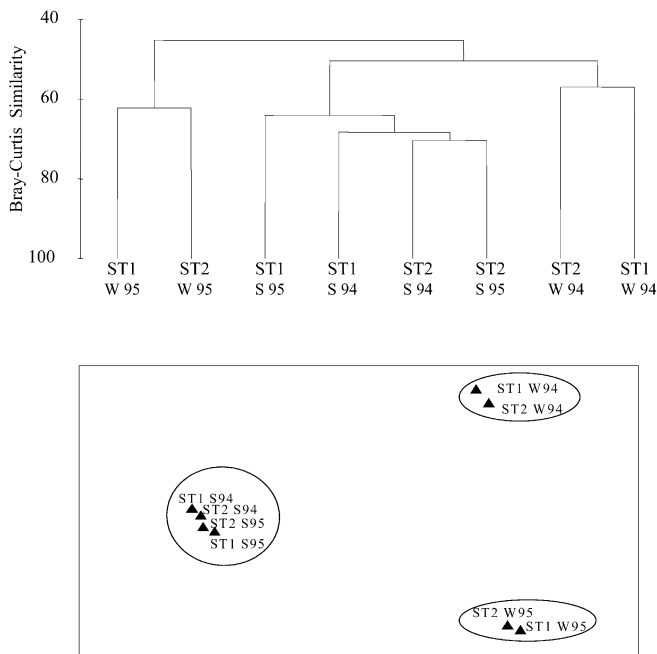
A total of 17,090 individuals, representing 100 faunal species, were examined. As shown in Table 1, the number of species and the diversity indices (H' , d and J'), were determined for each sampling station and season (winter/summer). The number of species ranged from 37 to 49 in winter, and from 45 to 60 in summer. The diversity indices ranged proportionately as they appeared to be higher in the summer samples (Table 1).

Table 1 (continued)

| Species | ST1 W94 | | ST2 W94 | | ST1 W95 | | ST2 W95 | | ST1 S94 | | ST2 S94 | | ST1 S95 | | ST2 S95 | | Total | | | |
|-----------------------------------|---------|-------|----------------|-----|----------------|-------|----------------|-------|---------|----------------|---------|-------|----------------|--------|---------|----------------|-------|-------|----------------|------|
| | f | mAb | D _p | mAb | D _p | mAb | D _p | f | mAb | D _p | f | mAb | D _p | f | mAb | D _p | f | mAb | D _p | |
| <i>Nematoneis unicomis</i> | | | | | | | | | | | | | | | | | | | | |
| <i>Nereis caudata</i> | 33 | 17 | 0.15 | | | | | 33 | 8 | 0.09 | 66 | 33 | 0.09 | 33 | 8 | 0.05 | 100 | 42 | 0.19 | |
| <i>N. zonata</i> | | | | | | | | | | | | | | | | | | 33 | 25 | 0.11 |
| <i>Ophiotromus pallidus</i> | | | | | | | 66 | 92 | 1.01 | 100 | 333 | 0.94 | | 100 | 275 | 1.50 | 33 | 225 | 1.03 | |
| <i>Perimeris caudifera</i> | 33 | 8 | 0.07 | | | | 33 | 8 | 0.09 | | | | | | | | 33 | 8 | 0.04 | |
| <i>Phyllodoce rubiginosa</i> | 66 | 25 | 0.22 | 100 | 42 | 0.33 | | | | 66 | 50 | 0.14 | 33 | 17 | 0.06 | | 100 | 33 | 0.15 | |
| <i>Platyeris dumerilii</i> | 33 | 8 | 0.07 | 66 | 75 | 0.60 | 33 | 8 | 0.09 | 33 | 8 | 0.02 | | | | 66 | 42 | 0.19 | 29 | |
| <i>Polydora caeca</i> | 66 | 17 | 0.15 | 66 | 67 | 0.53 | | | | | 66 | 42 | 0.16 | 33 | 25 | 0.14 | 66 | 50 | 0.23 | |
| <i>P. ciliata</i> | 66 | 33 | 0.30 | | | | 33 | 8 | 0.09 | | 33 | 17 | 0.06 | 100 | 83 | 0.45 | 33 | 19 | 0.105 | |
| <i>Polyophthalmus pictus</i> | 33 | 8 | 0.07 | | | | | | | 33 | 17 | 0.06 | | | | | 8 | 3 | 0.018 | |
| <i>Potamilla reniformis</i> | | | | | | | | | | | | | | | | | | | | |
| <i>Potamoceros triquetus</i> | 100 | 283 | 2.52 | 33 | 33 | 0.27 | | | | 100 | 42 | 0.12 | | 33 | 8 | 0.05 | 33 | 46 | 0.257 | |
| <i>Priotonospio madagrensi</i> | | | | | | | 33 | 8 | 0.11 | | | | | 33 | 142 | 0.77 | 100 | 208 | 0.96 | |
| <i>Sabellaria spinulosa</i> | 66 | 25 | 0.22 | 33 | 150 | 1.20 | | | | 33 | 25 | 0.07 | 33 | 33 | 0.13 | 33 | 17 | 0.09 | 46 | |
| <i>Serpula concharum</i> | | | | | | | 33 | 158 | 1.26 | 33 | 25 | 0.07 | 33 | 433 | 1.63 | | 12 | 77 | 0.433 | |
| <i>S. vermicularis</i> | 33 | 17 | 0.15 | 100 | 150 | 1.20 | 66 | 92 | 1.22 | 100 | 508 | 1.44 | 66 | 1,575 | 5.93 | 100 | 283 | 1.54 | 79 | |
| <i>Spiridobranchus polyrema</i> | 66 | 33 | 0.30 | 100 | 2,483 | 19.79 | 33 | 8 | 0.11 | 100 | 533 | 1.51 | 100 | 150 | 0.56 | 100 | 267 | 1.45 | 21 | |
| <i>Staurocephalus rudolphii</i> | | | | | | | 33 | 8 | 0.11 | | | | | | | | 79 | 494 | 2,774 | |
| <i>Syllis krobni</i> | 33 | 8 | 0.07 | | | | 33 | 8 | 0.11 | 66 | 17 | 0.05 | 66 | 33 | 0.13 | 33 | 8 | 0.05 | 25 | |
| <i>S. prolifera</i> | | | | | | | 33 | 8 | 0.07 | 66 | 17 | 0.05 | 33 | 33 | 0.13 | 33 | 8 | 0.047 | 25 | |
| <i>Terebella lapidaria</i> | 33 | 8 | 0.07 | | | | 33 | 8 | 0.11 | 33 | 8 | 0.09 | 33 | 58 | 0.22 | 100 | 242 | 1.32 | 33 | |
| <i>Cerithium repes</i> | | | | | | | | | | | | | | | | | | | | |
| <i>Hanomia ephippium</i> | | | | | | | | | | 33 | 8 | 0.02 | | | | | | | | |
| <i>Hiatella rugosa</i> | 66 | 50 | 0.44 | | | | 33 | 8 | 0.09 | 33 | 8 | 0.02 | 66 | 208 | 0.78 | | 33 | 8 | 0.04 | |
| <i>Hexaplex trunculus</i> | | | | | | | | | | 66 | 125 | 0.35 | 100 | 417 | 1.57 | | 66 | 133 | 0.61 | |
| <i>Hinia incrasata</i> | 100 | 5,142 | 45.7 | 100 | 4,475 | 35.66 | 100 | 2,300 | 30.60 | 100 | 8,375 | 23.7 | 100 | 10,408 | 39.18 | 100 | 5,108 | 27.83 | 100 | |
| <i>Mytilus gallo-provincialis</i> | | | | | | | 33 | 8 | 0.07 | | | | | | | | | | | |
| <i>Odosmia sp.</i> | | | | | | | 33 | 8 | 0.11 | | | | | | | | | | | |
| <i>Aora sp.</i> | | | | | | | 33 | 8 | 0.07 | 33 | 8 | 0.11 | | | | | | | | |
| <i>Amphithoe ramondi</i> | 33 | 17 | 0.15 | | | | 33 | 8 | 0.07 | | | | | | | | | | | |
| <i>Athanas nitescens</i> | 100 | 233 | 2.07 | 66 | 125 | 1.00 | 66 | 50 | 0.67 | 33 | 8 | 0.09 | 66 | 250 | 0.94 | 100 | 200 | 1.09 | 79 | |
| <i>Balanus eburneus</i> | | | | | | | 100 | 100 | 1.33 | | | | | | | | 33 | 17 | 0.09 | |
| <i>B. perforatus</i> | 100 | 617 | 5.48 | 100 | 233 | 1.86 | 100 | 1,592 | 21.18 | 100 | 183 | 0.52 | 100 | 233 | 0.88 | 100 | 208 | 1.13 | 21 | |
| <i>B. trigonus</i> | 66 | 167 | 1.48 | 100 | 1,383 | 11.02 | 100 | 775 | 10.31 | 100 | 1,092 | 3.09 | 100 | 717 | 2.70 | 100 | 2,600 | 14.16 | 100 | |
| <i>Copepoda</i> | 100 | 117 | 1.04 | 66 | 100 | 0.80 | 33 | 8 | 0.11 | 33 | 125 | 0.35 | 33 | 58 | 0.22 | 100 | 917 | 4.21 | 100 | |
| <i>Corophium acuteratum</i> | | | | | | | | | | 100 | 592 | 1.67 | 33 | 8 | 0.03 | | | | | |
| <i>C. acutum</i> | | | | | | | 66 | 50 | 0.67 | 66 | 308 | 0.87 | 66 | 42 | 0.16 | 100 | 83 | 0.45 | 54 | |
| <i>C. sextonae</i> | 33 | 8 | 0.07 | 66 | 17 | 0.13 | 33 | 8 | 0.11 | 66 | 25 | 0.07 | 66 | 50 | 0.55 | 66 | 25 | 0.26 | 33 | |
| <i>Corophium sp.</i> | | | | | | | 33 | 25 | 0.28 | 100 | 11,400 | 32.26 | 100 | 6,583 | 24.78 | 100 | 4,842 | 26.37 | 54 | |
| <i>Erichthonius brasiliensis</i> | | | | | | | 33 | 8 | 0.09 | 33 | 17 | 0.05 | 66 | 100 | 0.38 | 100 | 66 | 42 | 0.23 | |

Table 2 Distribution of the species found in the assemblage to the level of major taxa

| Taxa | Number of species | Percentage % |
|---------------|-------------------|--------------|
| Polychaeta | 45 | 37.5 |
| Crustacea | 31 | 30.39 |
| Mollusca | 9 | 8.82 |
| Turbellaria | 4 | 3.92 |
| Cnidaria | 3 | 2.94 |
| Bryozoa | 2 | 1.96 |
| Echinodermata | 2 | 1.96 |
| Nemertina | 1 | 0.98 |
| Nematoda | 1 | 0.98 |
| Sipunculida | 1 | 0.98 |
| Pantopoda | 1 | 0.98 |
| Ascidiacea | 1 | 0.98 |
| Pisces | 1 | 0.98 |

**Fig. 2** Affinity of the sampling sites according to cluster and non-metric multidimensional sampling analyses

Abundance

The comparison of the faunal abundance in time (within and between years) and in space (sampling sites) was based on the examination of the null hypothesis that the abundance of the fauna does not differ significantly. One-way ANOVA was used to detect the exact differences and showed an equal distribution of the abundance of the fauna in winter as well as in summer samples ($F=0.58$, $df=1,3$, $P=0.64$ in winter samples, and $F=2.47$, $df=1,3$, $P=0.14$ in summer samples).

Affinity of sampling in space and time

The affinity of all sampling sites is given in Fig. 2. Both analyses (cluster and non-metric mds) indicate a separation of samples into three main groups at about 50% similarity level. The winter samples from 1995 form group A, the winter samples from 1994 group B and the summer samples from both years (1994 and 1995) group C. Summer and winter samples are separated, indicating seasonality of the benthic assemblages. The two stations were not separated. Samples from successive years were only separated in winter. The stress value for the two-dimensional mds configuration is 0.01, indicating an excellent ordination of samples (Clarke and Warwick 1994). The performance of a one-way ANOSIM test gave global $R=1$ at a significance level of $P<0.005$, so the separation of the three groups (A, B and C) was confirmed. Group A showed an average similarity of 62%. As identified by SIMPER analysis, five species (*M. galloprovincialis*, *B. perforatus*, *Elasmopus rapax*, *Pisidia longicornis* and *B. trigonus*) were responsible for 60% of the average similarity, and 15 species for 90%. Group B reached an average similarity of 57%, with seven species (*M. galloprovincialis*, *Pisidia longicornis*, *Ophiothrix fragilis*, *B. perforatus*, *B. trigonus*, *Pilumnus hirtelus* and *Athanas nitescens*) covering 60% of this similarity, and 22 species 90%. Finally, group C reached an average similarity of 67%, with eight species (*Corophium* sp., *M. galloprovincialis*, *B. trigonus*, *O. fragilis*, *Pisidia longicornis*, *Prionospio malmgreni*, *S. vermicularis* and *Pilumnus hirtelus*) being responsible for 60% of the similarity, and 26 species for 90%. As regards the divergence between groups, we found that group A had an average dissimilarity percentage of 50% with group B (21 species contributed 60% and 55 species 90% of this value) and 57% with group C (16 species contributed 60% and 54 species 90%), while the dissimilarity percentage between groups B and C was 50% (22 species contributed 60%, and 58 species 90%). It is quite clear that only a few species are important for characterizing the groups, while most of them are important for differentiating the groups.

Discussion

Photophilic soft algae communities probably have the highest faunal and flora diversity of all benthic biocoenoses in the Mediterranean (Pérès 1982; Bellan-Santini et al. 1994). The assemblage of mussel beds of *M. galloprovincialis* in Thermaikos Gulf belongs to this type of biocoenosis, showing high diversity with respect to species abundance. One hundred animal species were recorded during this study, 17 of which were characterized as very common ($f \geq 50\%$), and 57 as common ($10\% < f < 50\%$). The recorded species have been reported by many authors as members of the assemblages of sublittoral photophilic algae (Bellan-Santini 1969; Saldanha 1974; Kocatas 1978; Bellan 1980; Hong 1983;

Table 3 Literature data on *M. galloprovincialis* assemblages. *Q* Quadrat, *C* core, *INFR* infralittoral, *MID* midlittoral, *P* polluted, *NP* non-polluted

| Source | Location | Ecological zone | Sampling method | Surface (cm ²) | No. faunal species | No. polychaete species | No. crustacean species | Pollution of biotope |
|------------------------------------|----------------------|-----------------|-----------------|----------------------------|--------------------|------------------------|------------------------|----------------------|
| Bellan-Santini (1969) | Marseilles | INFR | Q | 400 | 43 | 13 | 14 | P |
| Bellan (1969, 1980) | Marseilles | INFR | Q | 400 | 98 | 25 | 24 | NP |
| Kocatas (1978) | E Aegean Sea | INFR | Q | 400 | 111 | 35 | 32 | P |
| Bellan (1980) | NW Mediterranean Sea | INFR | Q | 400 | | 20/114.3 | | NP |
| Bellan (1980) | NW Mediterranean Sea | INFR | Q | 400 | | 20/78.2 | | P |
| Thiel and Ullrich (2002) | Chile | INFR | Q | 100 | 62 | 15 | 14 | NP |
| Svanne and Setyobudiandi (1996) | Denmark | INFR | C | 282 | 43 | 13 | 10 | NP |
| Lintas and Seed (1994) | N Wales | MID & INFR | Q | 25 | 59 | 4 | 25 | NP |
| Tiganus (1979) | Black Sea | INFR | Q | ?? | ?? | ? | ? | ? |
| Topaloglou and Kihara (1993) | Bosporus | INFR | Q | 400 | 48 | 10 | 22 | NP |
| D'Anna et al. (1985) | Sicily, Italy | INFR | Q | 400 | 89 | 32 | 8 | P & NP |
| Tsuchiya and Nishihira (1986) | N Japan | INFR | Q | 100 | 69 | 27 | 17 | NP |
| Tsuchiya and Bellan-Santini (1989) | Marseilles | INFR | Q | 100 | 99 | 33 | 30 | NP |
| Damianidis and Chintiroglou (2000) | Thermaikos Gulf | INFR | Q | 400 | | 48 | | NP |
| Saldanha (1974) | Portugal | INFR | Q | 500 | 131 | 27 | 36 | NP |
| Saldanha (1974) | Portugal | MID | Q | 500 | 67 | 10 | 8 | NP |

Desrosiers et al. 1986; Marinopoulos 1988; Chintiroglou and Koukouras 1992; Topaloglou and Kihara 1993; Bellan-Santini et al. 1994; Lantzouni et al. 1998; Baxevanis and Chintiroglou 2000; Damianidis and Chintiroglou 2000).

The biodiversity of *M. galloprovincialis* assemblages in Thermaikos Gulf differs from those reported for other Mediterranean and northern European sites (Table 3). Most authors have used similar methods for the study of hard substrates (based mostly on work by Bellan-Santini 1969) (Table 3), yet data from different geographical areas are difficult to compare. The highest species abundance ($R=131$) was found in the infralittoral zone of the Portuguese coast. The clean waters of Marseilles, Ismir and Thermaikos Gulf also show high species abundance, while species abundance was relatively low in the Bosporus, on Danish coasts, in the midlittoral zone of the Portuguese coast and in the polluted waters of Marseilles. The respective values in Japan and Chile ranged around 65, while 89 species were counted in Italy and 56 in Aberffraw (N Wales).

Table 3 also shows that polychaetes and crustaceans are the most important taxonomic groups of the assemblage, contributing almost 50% of the total faunal species abundance. The abundance of these groups, however, varies from one area to another and also seems to be dependent on the specific features of each study area (polluted/non-polluted; midlittoral/supralittoral), a fact that was also reported by Thiel and Ullrich (2002). Saldanha (1974), for instance, recorded fewer polychaetes and more crustacean species on the coasts of Portugal,

while D'Anna et al. (1985) recorded exactly the opposite for Sicily. Kocatas (1978) found 35 polychaete and 32 crustacean species in the Izmir Bay, whereas Topaloglou and Kihara (1993) reported 10 polychaete and 22 crustacean species for the Bosporus. Seed and Suchanek (1992) drew the same conclusions, even though the Mediterranean mussel assemblages were not the focus of their work. The total diversity of mussel assemblages in the studied areas shows no significant variations. Furthermore, significant similarities exist at the taxonomic level. It should also be noted that, according to Damianidis and Chintiroglou (2000), there is no significant difference in the composition of the dominant species; this implies that the structure of the polychaete fauna of *M. galloprovincialis* assemblages in the Mediterranean Sea is largely homogeneous.

An important part of the study of mussel assemblages is the study of interactions between the structure of mussel populations and the associated fauna (Tsuchiya and Nishihira 1986; Lintas and Seed 1994). These studies have produced contradictory results. According to Tsuchiya and Nishihira (1986), the structure of mussel beds in the Pacific Ocean (Japan) has a direct effect on the diversity of the assemblages. Lintas and Seed (1994) suggested that the fauna associated with *M. edulis* appear to be related largely to mussel density. Damianidis and Chintiroglou (2000) reported similar results on the polychaete fauna of mussel beds in Thermaikos Gulf. In the present study, however, no relationship was found between mussel density and the abundance and diversity of the assemblage. As regards the structure (= distribution

of size classes of the populations) of the *M. galloprovincialis* populations, there is certain information in Le Breton and Chintiroglou (1998) indicating an uneven distribution in space and time. Although information is still limited, it appears that biotic interactions have a strong effect on the composition of the assemblage, which seems to decrease when the assemblage is studied as a whole. In this case, the composition of the assemblages is largely even in space and time. Nevertheless, attention should be paid to any variations in order to understand the biotic interactions in hard substrate assemblages (see Damianidis and Chintiroglou 2000).

The evenness of the studied assemblage showed some variation in space, and particularly in time, while the total homogeneity of the faunal composition remained unaffected. The separate analyses conducted by Lantzouni et al. (1998) and Damianidis and Chintiroglou (2000) produced similar results.

In conclusion, there is now adequate information about *M. galloprovincialis* assemblages, although there are still open questions. For biomonitoring of marine benthic assemblages, knowledge of the structure and function of *M. galloprovincialis* assemblages can play an important role. The main advantage of studying such assemblages, on a smaller or wider scale, is their similar physiognomic appearance (Reish 1971; Wenner 1988). As these assemblages can be found in clean as well as in polluted waters, they are of great interest in biomonitoring studies (see Wenner 1988).

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